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1 Amazon deforestation: A dangerous future indicated by patterns and 2 trajectories in a hotspot of forest destruction in Brazil

3
4 *Beatriz Figueiredo Cabral¹, Aurora Miho Yanai¹, Paulo Maurício Lima de Alencastro
5 Graça¹, Maria Isabel Sobral Escada², Cláudia Maria de Almeida², Philip Martin
6 Fearnside¹*

7
8 ¹Instituto Nacional de Pesquisas da Amazônia (INPA), Av. André Araújo n° 2936, CEP 69067-
9 375, Manaus, Amazonas, Brazil, beatriz.figueiredocabral@gmail.com; yanai@inpa.gov.br

10 pmlag@inpa.gov.br; pmfearn@inpa.gov.br

11 ²Instituto Nacional de Pesquisas Espaciais (INPE), Av. dos Astronautas, 1758, CEP 12227-010,
12 São José dos Campos, São Paulo, Brazil, isabel.escada@inpe.br; claudia.almeida@inpe.br

13 14 15 **ABSTRACT**

16 In recent years, the loss of forest in the Brazilian Amazon has taken on alarming
17 proportions, with 2021 recording the largest increase in 13 years, particularly in the
18 Abunã-Madeira Sustainable Development Reserve (SDR). This has significant
19 environmental, social, and economic repercussions globally and for the local
20 communities reliant on the forest. Analyzing deforestation patterns and trends aids in
21 comprehending the dynamics of occupation and deforestation within a critical Amazon
22 region, enabling the inference of potential occupation pathways. This understanding is
23 crucial for identifying deforestation expansion zones and shaping public policies to curb
24 deforestation. Decisions by the Brazilian government regarding landscape management
25 will have profound environmental implications. We conducted an analysis of
26 deforestation patterns and trends up to 2021 in the municipality (county) of Lábrea,
27 located in the southern portion of Amazonas state. Deforestation processes in this area
28 are likely to spread to the adjacent “Trans-Purus” region in western Amazonas, where
29 Amazonia’s largest block of remaining rainforest is at risk from planned highways.
30 Annual deforestation polygons from 2008 to 2021 were categorized based on
31 occupation typologies linked to various actors and processes defined for the region (e.g.,
32 diffuse, linear, fishbone, geometric, multidirectional, and consolidated). These patterns
33 were represented through 10 × 10 km grid cells. The findings revealed that Lábrea’s
34 territory is predominantly characterized by the diffuse pattern (initial occupation stage),
35 mainly concentrated in protected areas. Advanced occupation patterns (multidirectional
36 and consolidated) were the primary contributors to deforestation during this period.
37 Observed change trajectories included consolidation (30.8%) and expansion (19.6%) in
38 the southern portion of the municipality, particularly along the Boi and Jequitibá
39 secondary roads, providing access to large illegal landholdings. Additionally, non-
40 change trajectories (67%) featured initial occupation patterns near rivers and in
41 protected areas, likely linked to riverine and extractive communities. Tailoring measures
42 to control deforestation based on actor types and considering stages of occupation is
43 crucial. The techniques developed in this study provide a comprehensive approach for
44 Amazonia and other tropical regions.

45
46 **Keywords** – deforestation actors; occupation typologies; cattle-ranching Amazon
47 frontier; land-use change; land grabbing

48 **1. Introduction**

49 *1.1. Amazon deforestation's consequences for Brazil and the world*

50 The central role of Amazon deforestation in the global climate and biodiversity
51 crises, as well as its environmental and social impacts in the Amazon region, have made
52 the containment of this destruction a top priority for both the world and for the
53 Amazonian countries (Fearnside, 2021a). With over 40% of the world's remaining
54 tropical forests, Brazil is by far the most important country in terms of potential impacts
55 of tropical deforestation and degradation on global climate (Laurance et al., 2001). This
56 stems not only from the substantial year-to-year carbon emission, but also from the
57 potential emissions of the huge carbon stocks in the remaining forest biomass and in the
58 soil under the forest if released by forest degradation from increasing temperatures,
59 drought severity and forest fires linked to climate change (Barros and Fearnside, 2019;
60 Nogueira et al., 2015). Release of this carbon over a span of a few years would be a key
61 factor in pushing global climate past a tipping point beyond which human
62 countermeasures by cutting emissions would be incapable of stopping the further
63 advance of global warming driven by positive feedbacks in the Earth system (Fearnside,
64 2022).

65 The municipality (county) of Lábrea, the focus of our study, is located in the
66 southern portion of Brazil's state of Amazonas and is adjacent to the Trans-Purus
67 region, a vast area that has largely remained undisturbed due to the lack of accessibility
68 by roads. The Trans-Purus region, located in the western portion of the state of
69 Amazonas, has an extensive area of "undesignated public forests," a term referring to
70 government land that has not been assigned to a specific use such as a conservation unit
71 (a protected area for biodiversity), an Indigenous land or a settlement project. These
72 undesignated public forests are the most attractive areas for illegal land occupation by
73 land grabbers, organized landless farmers, loggers, and other groups. The occupation
74 processes occurring in Lábrea are likely to expand to the Trans-Purus region (Fearnside,
75 2017, 2018; Fearnside et al., 2020). The Trans-Purus region faces deforestation risks
76 from roads planned to branch off the BR-319 (Manaus-Porto Velho) highway. BR-319
77 highway was initially built in 1972-1973 and abandoned in 1988; it received
78 "maintenance" beginning in 2015 and is now slated for "reconstruction," pending
79 environmental license approval (Figure 1) (Fearnside, 2022; Santos et al., 2023).

80 If the forest in the Trans-Purus region is lost it would be catastrophic for Brazil,
81 since this area is crucial for recycling the water that is transported to São Paulo and
82 other parts of southeastern and southern Brazil via the winds known as "flying rivers"
83 (Fearnside, 2004, 2015). The prevailing winds in Amazonia blow from east to west
84 because of the effect of the Earth's rotation on the return flow of the Hadley circulation,
85 and these winds bring water that has evaporated from the Atlantic Ocean. Roughly half
86 of the water that falls as rain over Amazon forest is recycled by the trees, which release
87 it as water vapor that, along with the remaining water vapor derived directly from the
88 Atlantic, is carried further by winds (Zemp et al., 2014). South American low-level jet
89 winds are blocked by the Andes mountains and turn to the south and east, transporting
90 water vapor to São Paulo, other parts of southeast and southern Brazil and to
91 neighboring countries (Nicolini et al., 2002; Arraut et al., 2012). This transport is
92 especially strong during December, January and February when the intertropical
93 convergence zone is at its southernmost position and the winds meet the highest part of
94 the Andes. This is rainy season in southeastern South America and is the critical period
95 when the reservoirs that supply the city of São Paulo fill, as well as the many other
96 reservoirs in southeastern and southern Brazil and in neighboring countries. The La
97 Plata River basin, which includes São Paulo, depends on water vapor from Amazonia

98 for 70% of its annual precipitation (van der Ent et al., 2010). A catastrophic drought in
99 2014 brought São Paulo, the fourth largest city in the world, close to running out of
100 water even for drinking (Nobre et al., 2016; Cunha et al., 2019; Fearnside, 2021b). In
101 2021 another catastrophic drought struck this part of Brazil (Fernandes et al., 2021;
102 Getirana et al., 2021). These droughts result from changes in ocean temperatures linked
103 to global warming, and they are expected to intensify with projected climate change
104 (Biaostoch and Boning, 2013; Coelho et al., 2016). These changes increase the value of
105 maintaining Amazon forest, especially in the Trans-Purus region, as there is no longer
106 any leeway to allow loss of water transported by the “flying rivers.”

107 *1.2. Amazon deforestation and the Lábrea hotspot*

108 From 2019 to 2021, annual deforestation in the Amazon biome in Brazil
109 increased by 56.6% when compared to the years from 2016 to 2018, surpassing 10,000
110 km² per year (INPE, 2021; Alencar et al., 2022). In 2021, the state of Amazonas jumped
111 to second place in the ranking for annual deforestation among the nine states in Brazil’s
112 Legal Amazon region (an area the size of western Europe), taking over the historical
113 position previously occupied by Mato Grosso and remaining behind only the state of
114 Pará (InfoAmazonia, 2021).

115 In recent years, deforestation has been concentrated mainly in eastern Acre,
116 northern Rondônia, and southern Amazonas, which make up the AMACRO integration
117 zone, or the recently renamed Abunã-Madeira Sustainable Development Zone.
118 AMACRO was created in 2021 during the 2019-2022 presidential administration of Jair
119 Bolsonaro with the justification of promoting environmental sustainability through the
120 economic development of the 32 municipalities that make up this region (SUDAM,
121 2021). However, the prospect of replacing the existing local economy with an
122 agribusiness-based economy aimed at exporting commodities is tied to the
123 intensification of rural violence, as observed in this area (CPT, 2021; IPAM, 2021;
124 IMAZON, 2022). As the potential for large-scale soy plantations and ranching
125 operations increases, there is heightened demand for land and an escalation of land
126 conflicts. These conflicts often manifest as “land grabbing,” a term specific to
127 Amazonia, denoting the large-scale, illegal claiming of land, typically on government-
128 owned territories (different from the same term’s meaning in Africa and Asia, where it
129 refers to foreign groups purchasing local land) (Agrawal et al., 2019). Government land
130 is invaded both by land grabbers and by small farmers (both organized and individual).
131 The result has been increased deforestation rates (IMAZON, 2022). The occupation
132 process in Lábrea reflects a historical context marked by intense land conflicts related to
133 illegal and disorderly occupation, expansion of cattle ranching, predatory fishing and
134 hunting, illegal logging, and landholdings in protected areas (Tavares and Cordeiro,
135 2017; CPT, 2021). The term “landholdings” used here refers to self-declared claims in
136 the Rural Environmental Registry (CAR) and does not imply that claimants have
137 ownership or legal title to the areas.

138 *1.3. Typology of deforestation patterns in the Amazon*

139 Addressing escalating deforestation and its profound consequences for humanity
140 is crucial, prompting the need for a comprehensive understanding of this dynamic
141 process. Given the intricate connection between deforestation and changes in land use, it
142 is crucial to comprehend and link the actors and factors involved, forming the basis for
143 targeted policies to mitigate the negative impacts of deforestation on the environment
144 and society (Marinaro et al., 2022). To achieve a thorough understanding of land
145 occupation, a robust approach involves discerning the typology of deforestation patterns
146 and correlating them with actors and different stages of occupation in a region
147 (McGarigal and Marks, 1995; Mertens and Lambin, 1997; Pereira et al., 2007; Silva et

148 al., 2008). This correlation could allow inferences about the evolutionary stage of the
149 occupation and deforestation process in a given area, including likely occupation
150 trajectories (Diniz, 2002; Oliveira and Metzger, 2006; Gavlak et al., 2011).

151 Previous studies in the Amazon have successfully employed typologies to
152 categorize deforestation patterns, shedding light on the characteristics associated with
153 various stages of occupation. For instance, Silva et al. (2008) identified distinct patterns
154 such as linear, small isolated irregular, small irregular near roads, medium irregular near
155 roads, and large geometric, each linked to specific actors and landholding
156 characteristics. Saito et al. (2011) classified deforestation based on occupation stages,
157 distinguishing between consolidated, diffuse, fishbone, regular geometric,
158 multidirectional disordered, and unidirectional linear patterns. Similarly, Gavlak et al.
159 (2011) defined deforestation patterns and trajectories along a specific highway,
160 providing insights into stages of occupation and associated changes. In a different study,
161 Marinaro et al. (2022) defined the stages and actors of deforestation in a livestock
162 frontier in the Dry Chaco of northern Argentina, contributing to the understanding,
163 organization, and differentiation of various social-ecological processes and perceptions
164 from local-urban inhabitants.

165 Building upon these insights, the region containing the municipality of Lábrea
166 presents an emergent scenario influenced by a cattle-ranching frontier, consolidating its
167 significance as an important area for comprehending land-occupation dynamics. To gain
168 insight into the recent dynamics of deforestation in the municipality of Lábrea, this
169 study endeavors to understand the patterns and trajectories of land occupation in the
170 region. Focusing on identifying deforestation patterns and trajectories in the
171 municipality of Lábrea from 2008 to 2021 will contribute valuable knowledge to inform
172 policies aimed at mitigating deforestation rates in this critical region.

173 **2. Materials and methods**

174 *2.1. Study Area*

175 The present study was carried out in the municipality of Lábrea, located in the
176 southern mesoregion of the state of Amazonas near the borders with the states of
177 Rondônia and Acre (Figure 1). The main rivers in the municipality are the Ituxi and
178 Purus, with the Purus River being used as the access route to the municipal seat (the city
179 of Lábrea; 2021 urban population = 24,223) (Tavares and Cordeiro, 2017; IBGE,
180 2021a). In addition to the river, it is also possible to access the city of Lábrea by the
181 BR-230 (Transamazon) highway, and the southern portion of the municipality can be
182 accessed by secondary roads branching off the BR-364 (Porto Velho-Rio Branco)
183 highway; these secondary roads, built illegally by landgrabbers, have played an
184 important role in the process of occupying this area (Franco, 2011; Tavares and
185 Cordeiro, 2017).

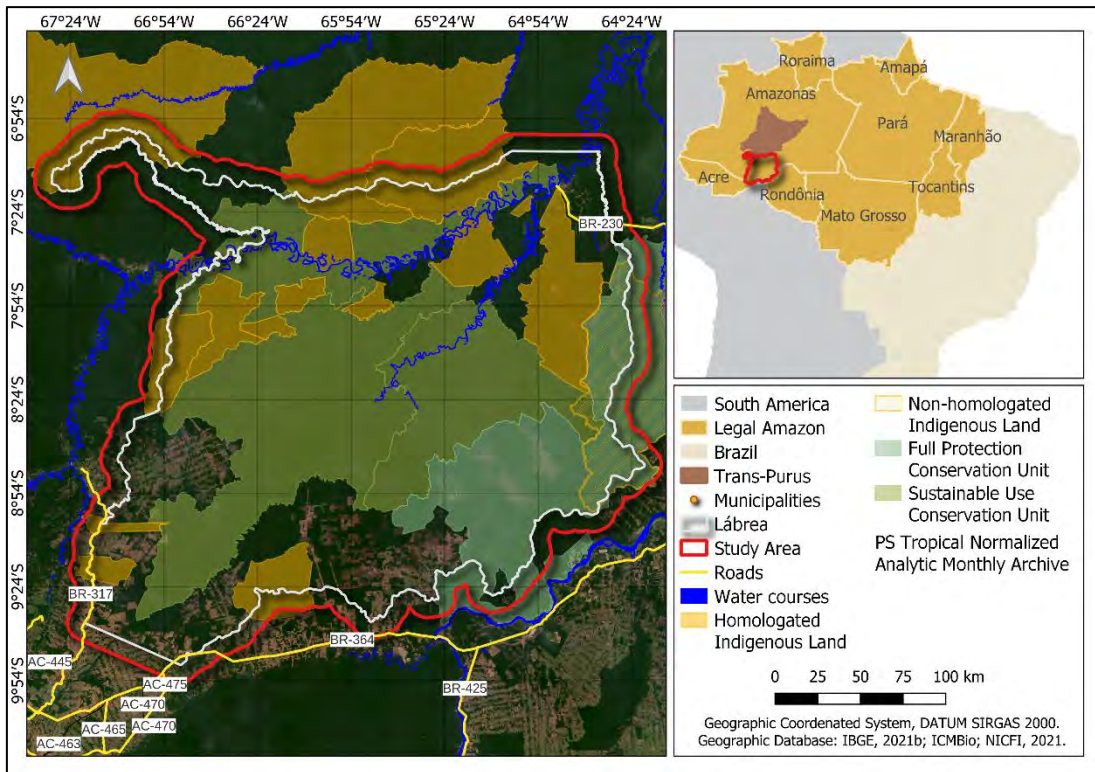


Figure 1. Location map of study area, municipality of Lábrea, Amazonas state.

The study area encompasses 83,900 km² (an area larger than Austria or the Brazilian state of Paraíba). The study area includes a 10-km buffer surrounding the municipality of Lábrea. The buffer was delineated to avoid possible information losses regarding deforestation patterns caused by the neighborhood of the municipality. Around 80% of the study area is in protected areas, divided into 17 Indigenous lands and four conservation units. The vegetation cover in Lábrea is predominantly composed of dense ombrophilous lowland forest (RADAMBRASIL, 1978). The soils are classified as Latosols, Argisols, Plinthosols, and, in areas near the rivers, Gleysols (IBGE, 2020). The relief is formed by depressions and plains, with altitudes ranging from 38 m to 1089 m above mean sea level. The climate is characterized as tropical monsoon (Type Am-i in the Köppen classification system), with short dry periods and a long period with high precipitation (2400 to 2800 mm year⁻¹) (Alvares et al., 2013).

2.2. Deforestation mapping and definition of occupation typology

Cumulative deforestation data for 2008, 2013, 2017, and 2021 were obtained from the Project for Monitoring Deforestation in the Legal Amazon by Satellite (PRODES) vector maps (INPE, 2021). The selection of these specific years aimed to analyze the distinct occupation patterns during periods with varying occupation dynamics. PRODES, at the National Institute for Space Research (INPE), estimates annual rates of deforestation by clearcutting based on visual interpretation of satellite images and provides extensive spatial data through the TerraBrasilis digital platform (Assis et al., 2019; INPE, 2021). Widely recognized for its accuracy, PRODES boasts a precision level of 93.5%, establishing its reputation as a reliable deforestation monitoring system (Maurano et al., 2019).

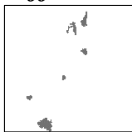
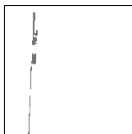




The analysis of deforestation patterns and trajectories using PRODES data proves invaluable in indicating and characterizing diverse processes of forest-cover change associated with agents and types of occupation, enabling a comprehensive

216 understanding of the deforestation and the occupation processes involved. This
217 approach involves creating a typology that establishes and describes the connection
218 between the spatial deforestation patterns in the analyzed area and their semantics.
219 Auxiliary information, such as agribusiness census data, and field information are used
220 in this process. Through exploratory spatial analysis of the arrangement and shape of
221 deforestation polygons and patterns described in the literature, an occupation typology
222 was defined for the Lábrea region based on Azeredo et al. (2016), Gavlak et al. (2011),
223 Saito et al. (2011), and Yanai et al. (2020). The identification and definition of these
224 patterns were also informed by an expedition carried out in the study area in October
225 2021, revealing different occupation dynamics.

226 After defining the typology of occupation patterns, the deforestation polygons
227 were analyzed in the context of individual cells. This process involved examining the
228 landscape metrics of the polygons within predetermined cell sizes. Previous studies
229 suggest that the choice of the grid-cell size should consider the dimensions of
230 deforestation patches and their spatial distribution to avoid loss of information or
231 distortion of patterns (Saito, 2010; Gavlak et al., 2011; Saito et al., 2012; Pinheiro et al.,
232 2016; Assis et al., 2021). After preliminary tests and an empirical analysis of the
233 arrangement of deforestation patterns when displayed with different cell sizes, it was
234 found that 10×10 km was the most suitable cell size for identifying and describing the
235 patterns proposed in Table 1. Cells where no deforestation patches were observed were
236 classified as "forest."
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Table 1. Typology of deforestation patterns associated with human occupation in the municipality of Lábrea (white = forest, gray = deforestation).

Deforestation Pattern	Description
<p><i>Diffuse</i></p> 	<p>Small, isolated patches with low to medium density, evenly distributed. This is characteristic of the initial spontaneous occupation distributed along riverbanks, in mining areas, and where small landholders are located.</p>
<p><i>Linear</i></p> 	<p>Elongated, continuous patches. This is related to the initial occupation distributed along roads.</p>
<p><i>Geometric</i></p> 	<p>Small and medium patches with regular geometry. This represents an initial to intermediate stage of occupation by medium and large landholders.</p>
<p><i>Fishbone</i></p> 	<p>Long, branching patches along roads, with an appearance similar to the skeleton of a fish. They represent an intermediate stage of occupation by small and medium landholders, as well as areas intended for settlement projects.</p>
<p><i>Multidirectional</i></p> 	<p>Large and medium-sized multidirectional patches with irregular geometry. They represent an intermediate to advanced stage of disorderly occupation.</p>
<p><i>Consolidated</i></p> 	<p>Large areas with undefined shape. They have a high density of deforestation with few or almost no forest fragments.</p>

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2.3. Classification of deforestation patterns and trajectories

Deforestation patterns were classified using spatial data-mining techniques. This procedure was performed using the Geographical Data Mining Analyst (GeoDMA) plugin, a classifier developed by Körting et al. (2008) for TerraView version 5.6.1 software. It combines a set of tools for integrating and analyzing remote-sensing images with data-mining techniques for extracting information and discovering knowledge about large databases (INPE, 2020). This tool allows for the structural classification of objects by extracting a set of landscape metrics that measure and describe the occupation patterns.

The classification process conducted by GeoDMA involved the following steps: 1) extraction of landscape metrics from deforestation polygons within 10×10 -km cells; 2) compilation of a comprehensive set of training samples for the deforestation class; 3) application of the C5.0 decision tree algorithm to classify and validate the deforestation

255 patterns (Quinlan, 2017), and 4) evaluation of the classification results by analyzing the
256 confusion matrix generated for the training samples.

257 During the training phase, a single set of samples was constructed for the four
258 analyzed years (i.e., 2008, 2013, 2017, and 2021). This allowed for patterns that are not
259 present in a given year to be included in another period, composing a unique typology
260 that considered deforestation patterns for all years. In total, 383 training samples were
261 collected, of which 60 samples corresponded to diffuse, forest, and geometric patterns,
262 58 samples were collected for the multidirectional pattern, 55 samples for the
263 consolidated pattern, and 45 samples for the fishbone and linear patterns. Due to the low
264 frequency of some deforestation patterns in the study area, such as the linear and
265 fishbone patterns, different numbers of samples were selected for each class.

266 Once the training samples were collected, we used the C5.0 algorithm with its
267 “boosting” feature, which generates a predefined number of trees where the final
268 classification of each segment is the one that was assigned by the majority of the
269 decision trees (Quinlan, 2017). GeoDMA automatically separated 66% (254 samples) of
270 the collected samples to classify the patterns (i.e., decision-tree generation), and 34%
271 (129 samples) for validation.

272 As a result of using the boosting tool, a predetermined set of 30 decision trees
273 was generated, where the categorical attributes (i.e., deforestation patterns) were
274 classified from the non-categorical attributes (i.e., values of landscape metrics). In total,
275 17 landscape metrics were used, of which three were in over 90% of the classifications,
276 namely: Class Area (CA), Landscape Shape Index (LSI), and Patch Richness (PR). The
277 following metrics also had a certain degree of importance (i.e., utilization values above
278 60%): Area-Weighted Mean Patch Fractal Dimension (AWMPFD), Total Area of the
279 Biggest Object (TABO), Patch Density (PD), Biggest Intersection Area (BIA), and
280 Area-Weighted Mean Shape Index (AWMSI). These metrics, by quantifying the spatial
281 structure and composition of landscapes, provide insights into the characteristics of
282 deforestation polygons in the grid, facilitating the identification of patterns, changes,
283 and trends over time. For a comprehensive understanding of these metrics, refer to the
284 Supplementary Material (Table S.1).

285 In the final step, two confusion matrices were generated to evaluate the result of
286 combining decision trees regarding the classification of training and validation samples
287 (Table 2). For the evaluation samples the total accuracy was 92% and the Kappa
288 coefficient was 91%, these values being in the range considered to be quite satisfactory
289 (Hudson and Ramm, 1987).

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Table 2. Confusion matrix of validation samples used in the classification of deforestation patterns for the study area for the four analyzed years (2008, 2013, 2017, and 2021) *

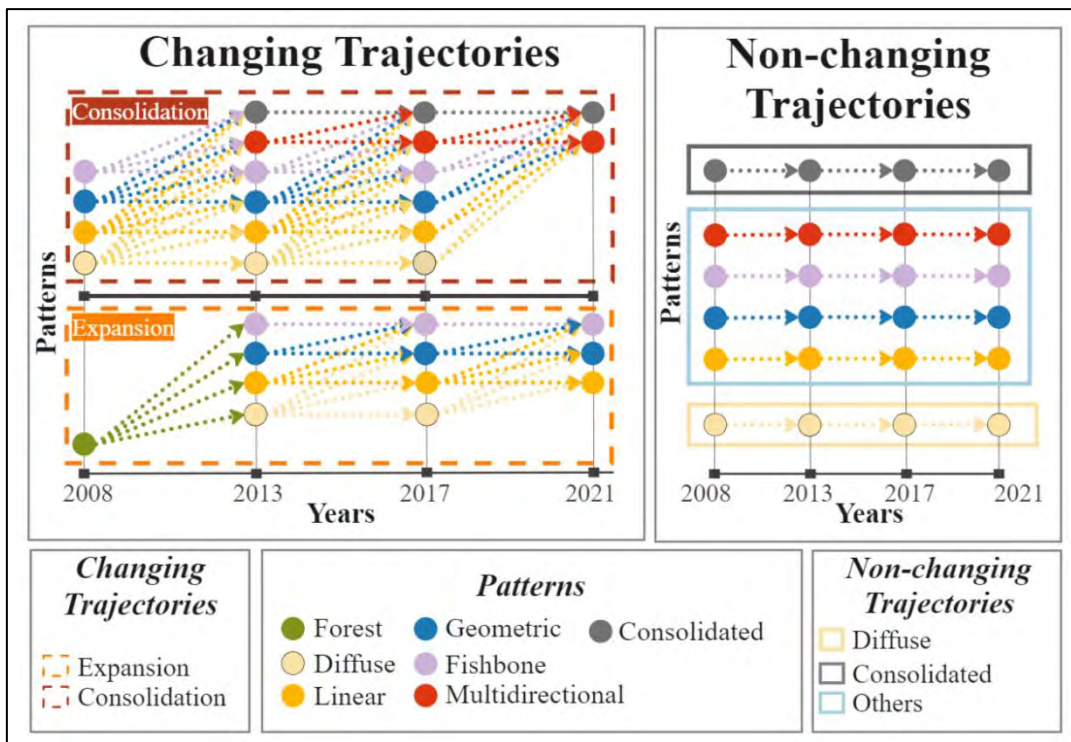
	CON.	DI.	FISH.	FO.	GE.	LI.	MUL.	P (%)
CON.	14							100
DI.		20				3		87
FISH.			15				1	94
FO.				20				100
GE.					24		1	96
LI.		1	2		1	13		76
MUL.			1				13	93
U (%)	100	95	83	100	96	81	87	
OA								0.92

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* The analyses use 129 validation samples. Con.= Consolidated; Dif. = Diffuse; Fish. = Fishbone; Fo. = Forest; Ge. = Geometric; Li. = Linear; Mul. = Multidirectional; P = Producer's Accuracy; U = User's Accuracy; OA: = Overall Accuracy.

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Deforestation trajectories between 2008 and 2021 were obtained from the results of the pattern classification. These trajectories were divided into "changing" and "non-changing," as proposed by Gavlak (2011). Change in trajectories refers to cells whose patterns were altered over the years, divided into "expansion" and "consolidated." The "expansion" trajectory is characterized by the cells initially having no deforestation (i.e., they were forest), but in the following years they were classified with patterns with initial to intermediate deforestation levels (i.e., diffuse, linear, geometric and fishbone). "Consolidation" is represented by all cells that in the first year presented an initial or intermediate occupation and by the last year had advanced to a multidirectional or a consolidated pattern (Figure 2). Cells that maintained the same pattern as in 2008 in the final year were classified as one of the "non-changing" trajectories according to the observed deforestation pattern, namely: diffuse, consolidated, and "others" (which encompasses the multidirectional, fishbone, geometric and linear patterns).



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313 **Figure 2.** Trajectories of change: consolidation and expansion, and non-changing
314 trajectories: diffuse, consolidated and others (Adapted from Gavlak, 2011).
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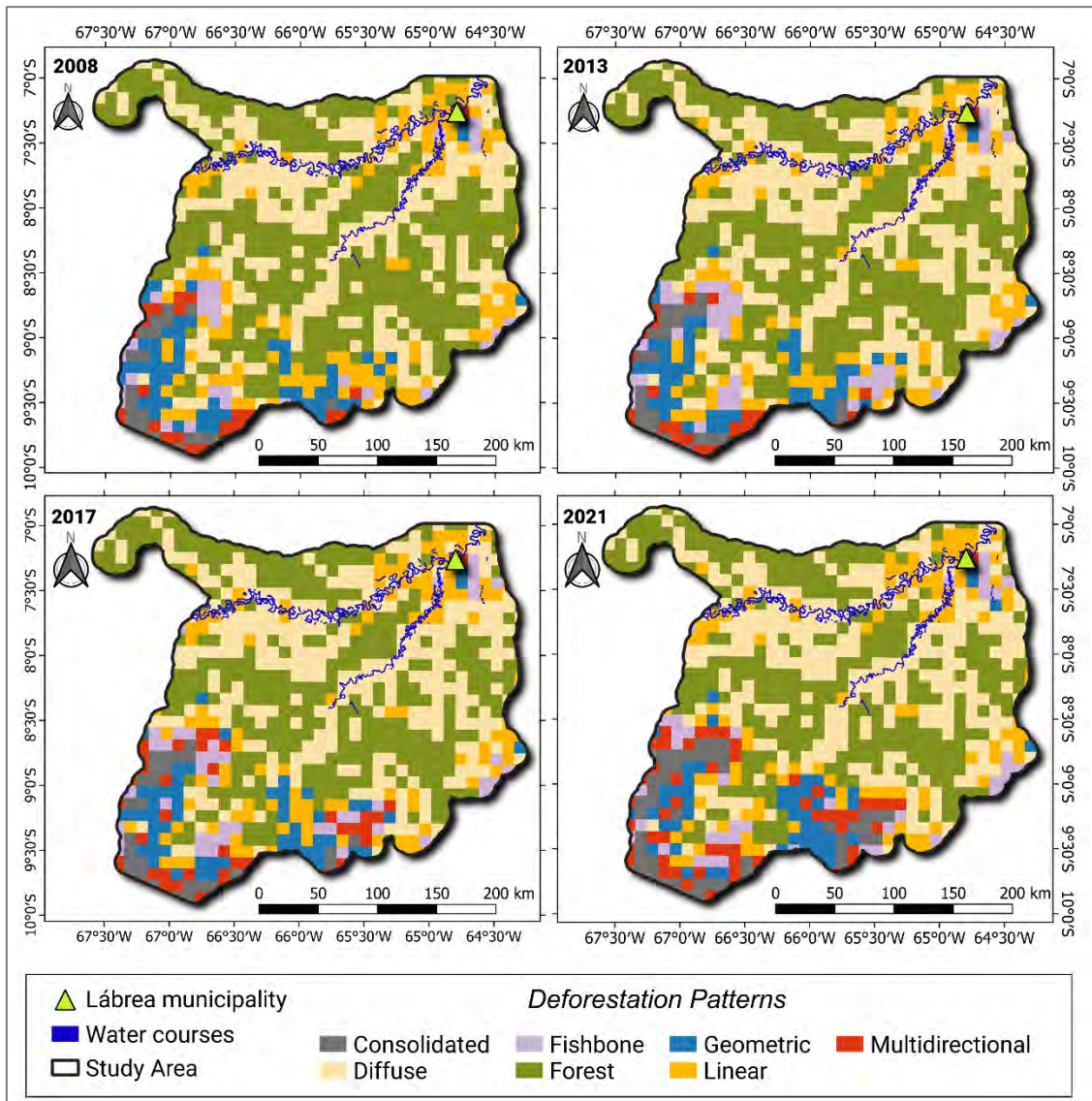
316 These trajectories were obtained both for change between the initial and final
317 maps (i.e., from 2008 to 2021), and in the three-time intervals between 2008 and 2021
318 (i.e., from 2008 to 2013, from 2013 to 2017, and from 2017 to 2021). The first analysis
319 made it possible to capture the differences in patterns for the entire period from 2008 to
320 2021, resulting in a single final map of occupation trajectories, while the second
321 allowed understanding the evolution of trajectories over the periods analyzed.

322 In this study the dataset does not allow for the identification of areas that have
323 reverted to forest once classified as deforested. This limitation reflects the relatively
324 short time frame of the analysis, which is insufficient to observe complete forest
325 regeneration.

326 2. Results

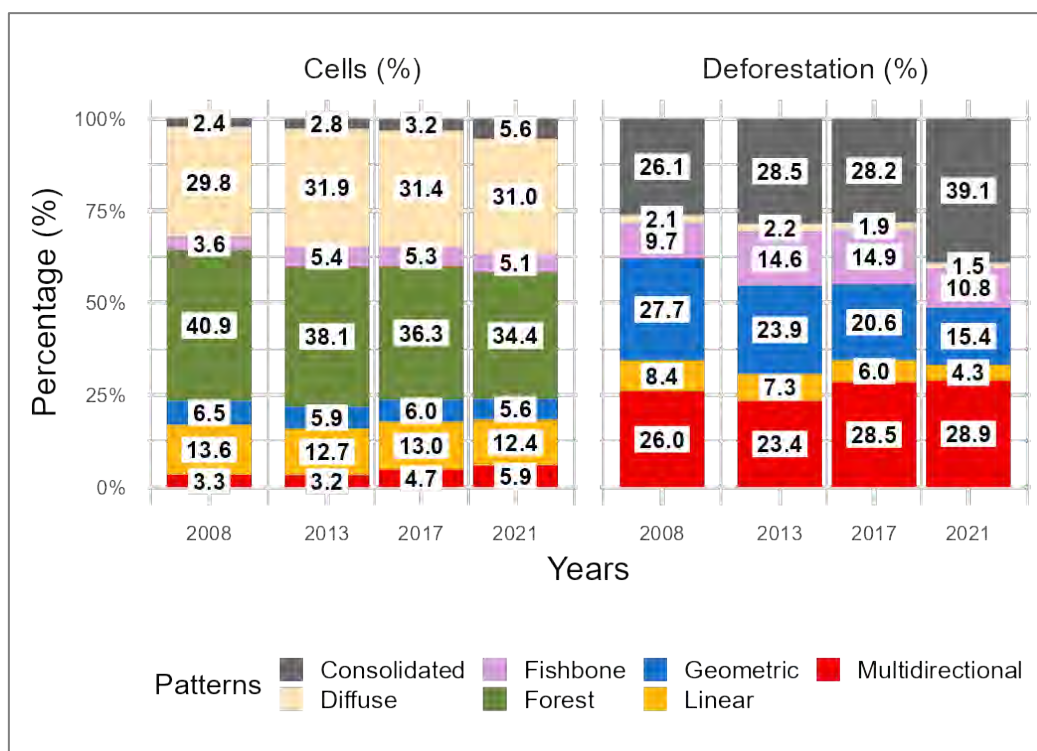
327 3.1. Deforestation patterns

328 The evolution of deforestation patterns from 2008 to 2021 is presented in the
329 maps in Figure 3. In general, it is possible to observe that the diffuse pattern (i.e., initial
330 occupation) was predominant and concentrated mostly in areas near rivers and in areas
331 that may indicate new deforestation fronts. Although the deforested area in cells
332 classified as diffuse increased over the years, from 12.7 to 14.6 thousand hectares from
333 2008 to 2021, as exemplified in Table 3, this amount observed in 2021 represented only
334 1.5% of the total deforested area, considering all patterns (Figure 4).
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Figure 3. Classification maps of occupancy patterns for the years 2008, 2013, 2017, and 2021 for the study area (i.e., municipality of Lábrea + 10-km buffer).



340
341 **Figure 4.** Evolution of the quantities of classified cells (%) and proportion of
342 cumulative deforested area (%) for each type of deforestation pattern for the years 2008,
343 2013, 2017, and 2021.

344
345 **Table 3.** Evolution of cumulative deforested area, in hectares, for each deforestation
346 pattern in the study area in 2008, 2013, 2017, and 2021.

Deforestation patterns	Cumulative deforested area (ha)			
	2008	2013	2017	2021
Consolidated	157,806.70	187,416.65	219,142.62	382,057.48
Diffuse	12,711.50	14,539.63	14,519.02	14,640.14
Fishbone	58,620.54	95,940.64	115,762.75	105,107.86
Geometric	167,178.21	157,325.14	159,579.75	150,823.57
Linear	50,554.12	48,251.45	46,211.81	41,737.13
Multidirectional	156,867.13	153,508.96	221,137.62	282,859.86
Total	603,738.19	656,982.46	776,353.58	977,226.04

347 The multidirectional and consolidated patterns, which represent advanced stages
348 of occupation, occurred mainly in the southern portion of the municipality of Lábrea.
349 Although these patterns had lower percentages of cells compared to the diffuse pattern,
350 they together represented 52.1%, 51.9%, 56.7%, and 68.0% of the deforested area for
351 the years 2008, 2013, 2017, and 2021, respectively (Figures 3 and 4). An analysis of the
352 deforested areas within these cells revealed a greater percentage increase in
353 deforestation for the consolidated pattern between 2017 and 2021. Over the course of
354 four years, the deforested area in consolidated cells increased by 162.9 thousand
355 hectares, whereas the multidirectional pattern showed an increase of only 61.7 thousand
356 hectares during the same period, and 67.6 thousand hectares from 2013 to 2017 (Table
357 3).

358 The linear and geometric patterns showed similar dynamics over the years
359 regarding the number of cells, with a reduction in area when comparing the years 2008
360 and 2021, reflecting a gradual reduction in cells classified as linear and geometric
361 (Figure 4). Since these patterns are related to the initial and intermediate stages of
362 occupation, respectively, the increase in deforested area in the cells implies a transition
363 to patterns in more advanced stages. Thus, over the years, a linear pattern cell changed
364 to the geometric or fishbone pattern, and the geometric pattern, in turn, progressed to
365 the multidirectional or consolidated pattern. Note that the fishbone pattern showed a
366 significant increase from 2008 to 2017.

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368 3.2. *Temporal evolution of deforestation trajectories*

369 The temporal evolution of trajectories during the three intervals between 2008
370 and 2021 (i.e., from 2008 to 2013, from 2013 to 2017, and from 2017 to 2021) are
371 illustrated in Figure S1. The change trajectories (i.e., “consolidation” and “expansion”)
372 and the “consolidated” no-change trajectory (i.e., cells that maintained the consolidated
373 pattern) presented similar frequencies with a slight increase over the years for the
374 consolidation and expansion trajectories but remained lower than the number of cells
375 with the diffuse and “others” trajectories (i.e., cells that maintained fishbone, linear, or
376 multidirectional patterns in the analyzed periods).

377 Both cells with the “consolidation” trajectory and cells with the “expansion”
378 trajectory showed increases in deforestation, and the cumulative deforested areas in both
379 classes increased over the years (Table 4). However, the contribution to deforestation in
380 Lábrea from areas undergoing consolidation was higher than in the expansion areas for
381 all three periods. The greatest increase in deforested area for both trajectories
382 (“consolidation” and “expansion”) occurred between the last two periods (i.e., between
383 2013 and 2017, and 2017 and 2021), with the cumulative deforestation from
384 consolidation increasing 2.7 times between 2013 and 2017, and 2.4 times from 2017 to
385 2021. The expansion trajectory had growth similar to these between 2013 and 2017
386 (increasing 2 times) and from 2017 to 2021 (increasing 1.6 times). This indicates that
387 the speed at which areas are undergoing consolidation is similar to areas undergoing
388 expansion.

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Table 4. Evolution of the cumulative deforested area, in hectares, for each deforestation trajectory and for the patterns that compose the "others" trajectories in the study area in the years 2013, 2017 and 2021.

Deforestation trajectory	Cumulative deforested area (ha)			Deforested area (%)		
	2013	2017	2021	2013	2017	2021
Consolidation	39,295.58	107,891.84	264,575.66	6.0%	13.9%	27.1%
Expansion	19,053.60	37,977.85	60,841.33	2.9%	4.9%	6.2%
Consolidated	160,245.84	192,254.68	229,299.02	24.4%	24.8%	23.5%
Diffuse	13,581.70	13,900.60	14,355.63	2.1%	1.8%	1.5%
Others						
Fishbone	82,004.08	95,643.96	90,573.58	-	-	-
Geometric	154,571.61	149,130.81	116,948.83	-	-	-
Linear	46,845.87	39,420.12	38,084.78	-	-	-
Multidirectional	141,384.17	140,133.72	162,547.21	-	-	-
Subtotal	424,805.73	424,328.62	408,154.40	64.6%	54.6%	41.7%
TOTAL	656,982.46	776,353.58	977,226.04	100%	100%	100%

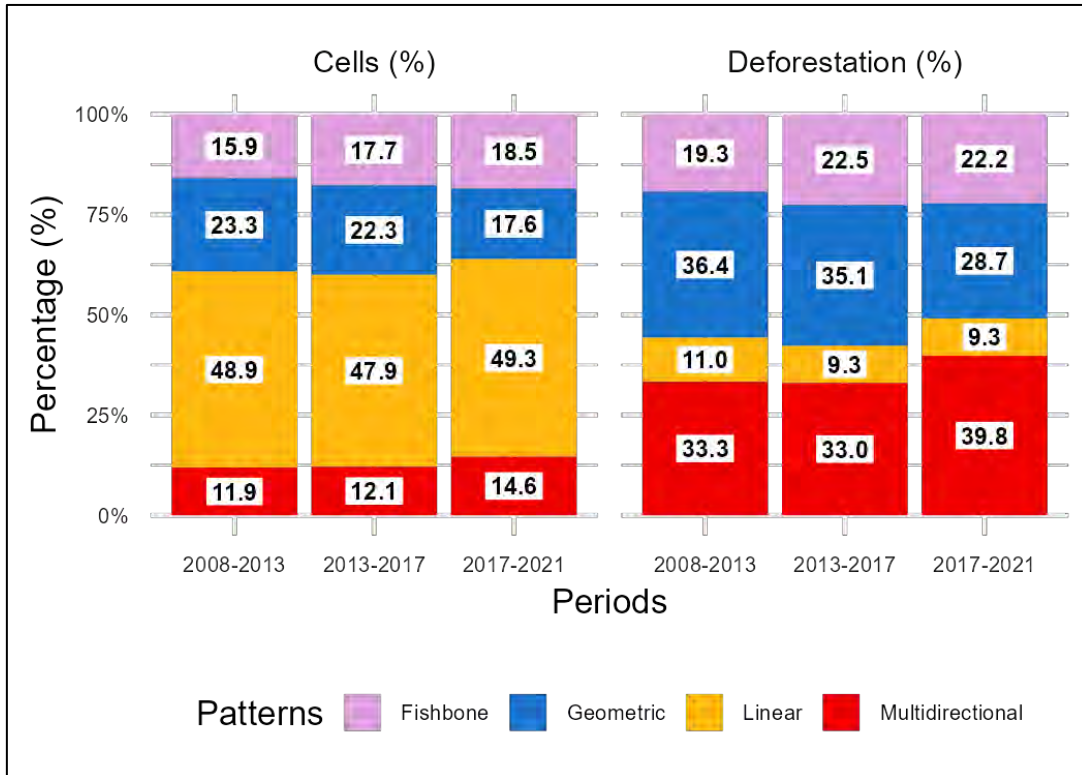
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The non-changing trajectories classified as diffuse and "others" showed similar behaviors in relation to cell frequency and the percentage of deforestation, decreasing from the initial value of 64.6% in 2013 to 41.7% in 2021 (Table 4). However, the relationship between the proportion of cells and the proportion of deforestation differed between these classes. Cells with the "others" trajectory accounted for the most deforestation in Lábrea (Table 4).

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The frequencies of patterns with their respective proportions of deforestation in cells classified in the "others" trajectory (i.e., proportion of non-change in the fishbone, geometric, linear and multidirectional patterns) are given in Figure 5. The geometric and multidirectional patterns had the highest contributions to the observed deforestation in this trajectory, so that the geometric pattern decreased while the percentage of cells and deforested area in the multidirectional pattern increased. This may indicate that deforestation increased in cells with the geometric pattern, causing them to no longer be part of the non-changing trajectory and changing to patterns of deforestation in more-advanced stages of occupation (i.e., there was an increase of deforestation in cells with geometric patterns that changed to multidirectional or consolidated patterns, noting that in the next analyzed period there was a smaller number of cells with the geometric pattern and a lower percentage of the deforestation represented by them).

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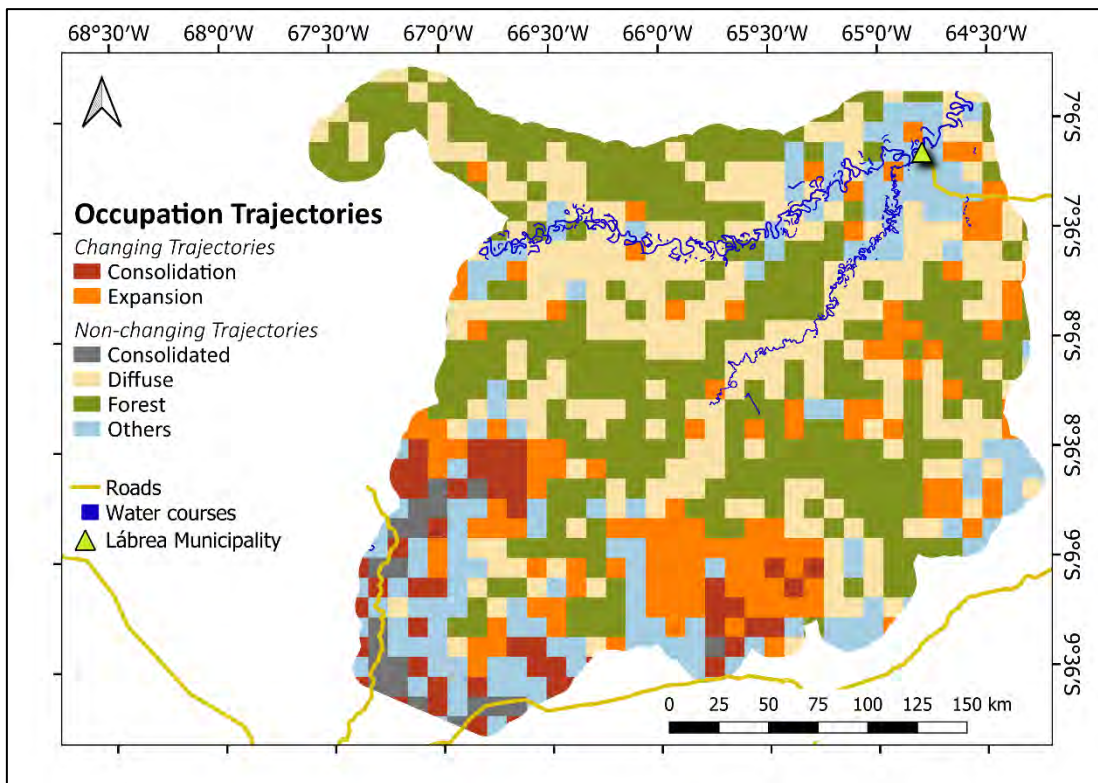
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Figure 5. Evolution of the occupation patterns that make up the non-change trajectories (i.e., “others”) in the study area in the periods from 2008 to 2013, 2013 to 2017, and 2017 to 2021.

For the linear and fishbone patterns, the proportions of deforested area varied from 9.3 to 22.5%, with distinct dynamics over time (Figure 5). An increase in deforested area was observed for the fishbone pattern, representing 22.2% of the total deforested area in the last period, which corresponds to 90.6 thousand hectares (Table 4). Cells that maintained the linear pattern, despite having a higher frequency than the previous pattern, had a reduction in deforested area over the years (Table 4). Both patterns are mainly related to early and intermediate stages of occupation, with changes to more advanced patterns being frequent over the years.

3.2. Deforestation trajectories

The final trajectory of deforestation for Lábrea (i.e., between the years 2008 and 2021), obtained from the analysis of pattern changes in different stages of occupation, is represented in Figure 6. Cells that presented change trajectories (i.e., “consolidation” and “expansion”) are mainly located in the southern portion of Lábrea, and, despite their low frequency compared to the others (29.4%), they represented 50.4% of deforestation in 2021 (Table S.2 in Supplementary material). The deforested area in cells undergoing consolidation and expansion is equivalent to 302,000 and 190,000 ha, respectively. Cells in expansion near rivers and in non-change trajectories with the diffuse pattern are mainly related to early occupations, as observed in Figures 3 and 4, and have smaller deforested areas.



438
439 **Figure 6.** Final map of deforestation trajectories from 2008 to 2021 for the study area
440 (i.e., the municipality of Lábrea + 10-km buffer). The “others” non-change trajectory is
441 represented by the fishbone, geometric, linear, and multidirectional patterns.
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443 The non-change trajectories are portrayed by cells that maintained their pattern,
444 namely "consolidated", "diffuse", and "others" (i.e., cells that maintained linear,
445 geometric, fishbone, and multidirectional patterns throughout the analysis period). The
446 "consolidated" non-change trajectory had the lowest percentage of cells (3.6%), but the
447 deforested area was 17.4% (Table S.2). These cells refer to areas where there was
448 intense deforestation prior to 2008 and were therefore already consolidated due to the
449 observed amount of deforestation. The total deforested area in these cells up to 2021
450 was 170,000 ha.

451 The highest frequencies observed belong to the diffuse and “others” non-change
452 trajectories (Figure 6). Cells that maintained the diffuse pattern (without change) over
453 the years had the highest frequency, representing 41.2% of Lábrea's total cells.
454 However, as previously mentioned, the diffuse pattern is characterized by areas of initial
455 occupation by riverine and extractivist communities with limited dynamics, so the
456 proportion of deforested area was only 1.4%, or 13,600 ha (Table S.2).

457 Cells classified as "others" comprised 25.8% of the study area (Table S.2). This
458 non-change trajectory was responsible for 30.8% of deforestation in Lábrea and
459 corresponded to an estimated area of 300,000 ha in 2021. Of the total cells classified as
460 "others," an area of 113,000 ha (11.6%) had the geometric deforestation pattern, and
461 96,000 ha (9.9%) had the multidirectional pattern.
462

463 **4. Discussion**

464 The patterns and trajectories of deforestation in Lábrea identified by the present
465 study have similarities in some respects to the occupation dynamics observed for this
466 region in previous studies. This is the case when comparing the results to a recent study
467 that used cellular grids with a resolution of 50×50 km for the entire Legal Amazon

468 region (Maurano et al., 2019). Similarities can be observed, such as intermediate and
469 advanced deforestation patterns (i.e., geometric and multidirectional, respectively)
470 concentrated in the southern portion of our study area, as well as the presence of the
471 diffuse pattern distributed throughout our study area. However, Maurano et al. (2019)
472 did not detect the consolidated pattern in Lábrea, indicating the loss of certain patterns
473 at coarser scales and, therefore, the importance of analyzing deforestation patterns and
474 trajectories at the municipal level to better understand local dynamics and how
475 deforestation frontiers advance in deforestation hotspots like Lábrea (Saito et al., 2011).

476 The "expansion" and "consolidation" trajectories were concentrated in the
477 southern portion of Lábrea, more specifically in two areas that were also reported by
478 Tomasi (2018): a portion near the Vista Alegre do Abunã district, which is crossed by
479 the secondary roads known as "Boi" and "Jequitibá," and another area near the
480 municipal seat of Boca do Acre, where the Monte Settlement Project (PA) is located
481 (Figures S.2-1 and S.2-2 in Supplementary material). In addition to these two areas,
482 small areas of expansion and consolidation were observed in our analysis near the
483 Mendes Junior secondary road (Figure S.2-3), located where the states of Amazonas,
484 Acre, and Rondônia meet. The intensification of this trajectory occurred mainly in the
485 last four years (2018-2021). The temporal evolution of deforestation patterns and
486 trajectories showed two important features: a large increase in the multidirectional and
487 geometric patterns (classified as an expansion trajectory in Figure 4) near the Boi and
488 Jequitibá secondary roads in 2021, and the presence of cells in consolidation where
489 deforestation is concentrated, followed by cells in expansion in the same area.

490 The dynamics observed in the study area underscore the role of roads as vectors
491 for the expansion of new deforested areas (Fearnside et al., 2009; Souza et al., 2017; dos
492 Santos Junior et al., 2018). Additionally, given the study area's proximity to the Tran-
493 Purus region, it is essential to highlight a few key insights. While the observed
494 evolution of patterns do not directly extend to the Tran-Purus region, it can still have
495 notable impacts (Figures 1 and 4). As is well known, the presence of protected areas
496 acts as the primary driver in reducing the advancement of deforestation into the northern
497 portion of the study area (Tomasi, 2018). However, this situation could change with the
498 degazetting of protected areas, environmental degradation, or the construction of new
499 roads leading to an escalating occupation process (Fearnside et al., 2009; Pack et al.,
500 2016). This last aspect is reinforced by three proposed projects presented by the
501 National Department of Transport Infrastructure (DNIT in Portuguese): the construction
502 of a state highway (AM-175) and two federal highways (BR-230 and BR-317) (DNIT,
503 2021a, 2021b; Figure S.3 in Supplementary material).

504 The geometric deforestation pattern is located near areas with advanced stages of
505 land occupation (i.e., multidirectional and consolidated) and is characteristic of
506 occupations by medium and large landholders, whose main activity is cattle ranching
507 (Gavlak et al., 2011; Saito et al., 2011). The regions in which these patterns are located
508 also coincide with the presence of claims in the Rural Environmental Registry (CAR)
509 that overlap with undesignated public lands (Figure S.4 in Supplementary material).
510 This type of overlap is an indication of land grabbing (Azevedo-Ramos et al., 2020;
511 IPAM, 2021; Moutinho, 2022).

512 Despite the presence of patterns that indicate intermediate and advanced stages
513 of occupation, the diffused deforestation pattern was predominant in Lábrea and was
514 concentrated mainly in the central part of the municipality, near the Purus and Ituxi
515 Rivers. In these areas, the presence of small landholdings and riverside populations with
516 low-impact agro-extractive practices is common, which characterizes the small
517 deforested polygons of the diffuse pattern (Schubart, 1983). Similar patterns were also

518 found by Gavlak (2011) in the Sustainable Forest District (DFS) of Highway BR-163 in
519 western Pará, and by Saito et al. (2011) in parts of the states of Pará, Mato Grosso, and
520 the southern edge of Amazonas.

521 In 2008, the federal government created four conservation units in the northern
522 portion of Lábrea, two extractive reserves (Resex Médio Purus and Ituxi), a national
523 park (PARNA Mapinguari) and a national forest (FLONA do Iquiri) (Figure 1). These
524 were created in the context of the proposed reconstruction of BR-319 connecting Porto
525 Velho to Manaus (Franco, 2011). Since there were no significant changes in the
526 frequency of the diffuse deforestation pattern over time, as well as the proportion of the
527 deforested area that this pattern represented, it is assumed that most of the deforestation
528 observed in the diffuse cells occurred before 2008. This underscores the impact of
529 protected areas in mitigating deforestation, even when they have a sustainable-use
530 designation, such as extractive reserves (areas designated for extraction of rubber,
531 Brazil nuts and other non-timber products), which inherently allow for more permissive
532 activities compared to areas under full protection. Despite this distinction, protected
533 areas play a crucial role in thwarting the progression of patterns associated with early-
534 stage occupations, preventing them from advancing into more advanced stages of
535 occupation and deforestation (Vitel et al., 2009; Soares-Filho et al., 2010; Walker et al.,
536 2020).

537 The linear pattern is associated with early stages of occupation mainly in the
538 vicinity of secondary roads. Depending on the region where they are located, cells with
539 this pattern can easily evolve into more advanced stages of deforestation, acting as a
540 vector for forest loss in previously inaccessible areas. In this study, the proportion of
541 cells classified as linear was small compared to other patterns. These cells were
542 distributed in areas near the municipal headquarters of Lábrea and in settlement
543 projects. It is important to note that the confusion matrices for the analyzed years
544 indicated classification errors between the linear pattern and the fishbone and geometric
545 patterns. These errors were also found by Gavlak (2011) due to the low frequency of
546 this pattern in his study area.

547 Although studies have related the fishbone pattern to the evolution of linear
548 patterns in some areas in the early stages of road construction, such as BR-319 and BR-
549 364 (Sampaio and Costa, 2009; Saito, 2010), this behavior was not so evident in the
550 present study. This may be associated with the stage of occupation in Lábrea, which, in
551 the analyzed years, presented more advanced occupation patterns (i.e., multidirectional
552 and consolidated). In addition, deforestation expansion from secondary roads connected
553 to main roads in the southern portion of the municipality occurred before 2008, with a
554 large portion of the trajectories observed being of the non-change type.

555 The evolution of patterns in existing settlement projects indicates that, until
556 2017, the Monte settlement project (located near the municipal seat of Boca do Acre)
557 displayed a fishbone pattern. By overlaying cells representing deforestation patterns
558 onto settlement boundaries, it became evident that the Gedeão Sustainable Development
559 Project (PDS) also exhibits fishbone patterns, despite being categorized as an
560 "environmentally differentiated" settlement project where actors and deforestation
561 patterns are expected to differ from a traditional settlement project (PA) (Yanai et al.,
562 2017). The presence of fishbone patterns in the initial analysis periods is linked to the
563 more intense occupation process in these settlement projects, particularly in the 2000s
564 (Tomasi, 2018).

565 Our findings reveal a significant intensification of deforestation in recent years
566 in the southern portion of Lábrea, specifically near the Boi and Jequitibá secondary
567 roads. This contrasts with the northern portion of the municipality, where the city of

568 Lábrea is located. The drivers behind the consolidation and expansion of deforestation
569 in this region are diverse, encompassing factors such as the dismantling of
570 environmental policies, increased conflicts in rural areas, weakened environmental
571 enforcement, land grabbing, illegal logging, and proximity to highways (Brito et al.,
572 2019; Ferrante and Fearnside, 2019; Azevedo-Ramos et al., 2020; Vale et al., 2021;
573 Wenzel, 2023).

574 Studies that identify the relationships between occupation dynamics and
575 economic, social, and environmental variables are crucial for defining strategic actions
576 to curb the deforestation that is advancing into forest areas in the municipality of
577 Lábrea. A recent study applying Bayesian techniques has shown that effectiveness of
578 past measures to slow Amazon deforestation depended heavily on the measures being
579 tailored to the types of actors present and to each area's degree of consolidation
580 (Brandão et al., 2023). Future studies should explore the relationship between the
581 observed deforestation patterns and fire occurrence, given their close association (Silva
582 et al., 2021; Dutra et al., 2023). Finally, the current study in the municipality of Lábrea
583 illustrates that these key variables and their trajectories can be quantified through
584 remote sensing. The application of these techniques could contribute to the
585 understanding of deforestation patterns, providing valuable insights for informing
586 political actions aimed at reducing deforestation rates not only within the eight-nation
587 PanAmazonian region, but also in other tropical forest areas globally.

588

589 **5. Conclusions**

590 The dynamics of deforestation in cattle-ranching frontier areas in the
591 municipality of Lábrea show that processes of occupancy and deforestation frontier
592 expansion do not occur uniformly. Identifying different trajectories of change is
593 essential for the development of actions to contain new illegal land occupations and the
594 increase in patterns that contribute most to deforestation, such as the multidirectional
595 and consolidated patterns that indicate advanced stages of occupation. It is also
596 important to focus efforts on areas with the consolidation trajectory, which, despite its
597 lower frequency, accounted for the largest deforested area in the municipality of Lábrea.

598 The areas bordering Lábrea in the neighboring the states of Rondônia and Acre
599 consisted mostly of areas with the consolidation trajectory, which had the highest
600 increase in forest loss over the 2017-2021 period. The growth of this trajectory is
601 mainly related to the increase in the multidirectional pattern (i.e., advanced stage of
602 occupancy). Next to these trajectories, expansion zones were identified that advance
603 from the southern part of Amazonas state into new areas to the north. This is
604 particularly concerning for the vast Trans-Purus region, located to the north and west of
605 Lábrea.

606 The patterns and trajectories of land-use change indicate how and where
607 deforestation occurs and allow identifying the types of actors responsible for this
608 dynamic. This technique can contribute to the development of public policies focused
609 on forest conservation and protection throughout Amazonia and in other tropical
610 regions. It is necessary to strengthen protected areas, which were observed to play an
611 important role in controlling deforestation. Effective measures to control deforestation
612 require tailoring them to the types of actors and to the degree of consolidation of
613 occupation. The techniques developed in this study can facilitate quantifying these
614 factors over wide areas in Amazonia and elsewhere in the tropics.

615

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626

627 **7. References**

- 628 Agrawal, A.; Brown, D.G.; Sullivan, J.A. 2019. Are global land grabs ticking socio-
 629 environmental bombs or just inefficient investments? *One Earth*, 1, 159–162.
 630 <https://doi.org/10.1016/J.ONEEAR.2019.10.004>.
- 631 Alencar, A.; Silvestrini, R.; Gomes, J.; Savian, G. 2022. Amazônia em chamadas – O
 632 novo e alarmante patamar do desmatamento na Amazônia. Nota técnica nº 9.
 633 Brasília, DF, Brazil: Instituto de Pesquisa Ambiental da Amazônia.
 634 (<https://bit.ly/3YmFdYG>). Accessed: 22 July 2022.
- 635 Alvares, C.A.; Stape, J.L.; Sentelhas, P.C.; Gonçalves, J.L.M.; Sparovek, G. 2013.
 636 Köppen's climate classification map for Brazil. *Meteorologische Zeitschrift*, 22(6),
 637 711-728. <https://doi.org/10.1127/0941-2948/2013/0507>.
- 638 Antunes, A.F.B.; Lingnau, C. 1997. Uso dos índices de acurácia para avaliação de
 639 mapas temáticos obtidos por meio de classificação digital. In: Anais do III
 640 Congresso e Feira para Usuário de Geoprocessamento. Curitiba, Paraná, Brazil:
 641 Sagres. <http://bit.ly/3Zmzqnh>.
- 642 Arraut, J.M.; Nobre, C.A.; Barbosa, H.M.; Obregon, G.; Marengo, J.A. 2012. Aerial
 643 rivers and lakes: Looking at large-scale moisture transport and its relation to
 644 Amazonia and to subtropical rainfall in South America. *Journal of Climate*, 25,
 645 543-556. <https://doi.org/10.1175/2011JCLI4189.1>.
- 646 Assis, L.F.F.G. Ferreira, K.R.; Vinhas, L.; Maurano, L.; Almeida, C.; Carvalho, A.;
 647 Rodrigues, J.; Maciel, A.; Camargo, C. 2019. TerraBrasilis: A spatial data
 648 analytics infrastructure for large-scale thematic mapping. *ISPRS International*
 649 *Journal of Geo-Information*, 8, art. 513. <https://doi.org/10.3390/ijgi8110513>
- 650 Assis, T.O.; Escada, M.I.S.; Amaral, S. 2021. Effects of deforestation over the Cerrado
 651 landscape: A study in the Bahia frontier. *Land*, 10(4), art. 352.
 652 <https://doi.org/10.3390/land10040352>.
- 653 Azeredo, M.; Monteiro, A.M.V.; Escada, M.I.S.; Ferreira, K.R.; Vinhas, L.; Pinheiro,
 654 T.F. 2016. Mineração de trajetórias de mudança de cobertura da terra em estudos
 655 de degradação florestal. *Revista Brasileira de Cartografia*, 68(4), 717–731.
 656 <https://doi.org/10.14393/rbcv68n4-44278>.
- 657 Azevedo-Ramos, C. Moutinho, P.; Arruda, V.L.S.; Stabile, M.C.C.; Alencar, A.; Castro,
 658 I.; Ribeiro, J.P. 2020. Lawless land in no man's land: The undesignated public
 659 forests in the Brazilian Amazon. *Land Use Policy*, 99, art. 104863.
 660 <https://doi.org/10.1016/j.landusepol.2020.104863>.
- 661 Barros, H.S.; Fearnside, P.M. 2019. Soil carbon is decreasing under “undisturbed”
 662 Amazonian forest. *Soil Science Society of America Journal*, 83(6), 1779-1785.
 663 <http://doi.org/10.2136/sssaj2018.12.0489>.
- 664 Biastoch, A.; Boning, C.W. 2013. Anthropogenic impact on Agulhas leakage.
 665 *Geophysical Research Letters*, 40, 1138–1143. <https://doi.org/10.1002/grl.50243>.

- 666 Brandão, F.; Befani, B.; Soares-Filho, J.; Rajão, R.; Garcia, E. 2023. How to halt
667 deforestation in the Amazon? A Bayesian process-tracing approach. *Land Use*
668 *Policy*, 133, art. 106866. <https://doi.org/10.1016/j.landusepol.2023.106866>.
- 669 Brito, B.; Barreto, P.; Brandão Jr., A.; Baima, S.; Gomes, P.H. 2019. Stimulus for land
670 grabbing and deforestation in the Brazilian Amazon. *Environmental Research*
671 *Letters*, 14(6), art. 064018. <https://doi.org/10.1088/1748-9326/ab1e24>
- 672 Coelho, C.A.S., de Oliveira, C.P., Ambrizzi, T. et al. 2016. The 2014 southeast Brazil
673 austral summer drought: regional scale mechanisms and teleconnections. *Climate*
674 *Dynamics*. 46, 3737–3752. <https://doi.org/10.1007/s00382-015-2800-1>.
- 675 CPT (Comissão Pastoral da Terra). 2021. Conflitos no Campo Brasil 2021: CEDOC
676 Dom Tomás Balduino – CPT. <http://bit.ly/3F0rtfo>.
- 677 Cunha, A.P.M.A.; Zeri, M.; Deusdará Leal, K.; Costa, L.; Cuartas, L.A.; Marengo, J.A.;
678 Tomasella, J.; Vieira, R.M.; Barbosa, A.A.; Cunningham, C. et al. 2019. Extreme
679 drought events over Brazil from 2011 to 2019. *Atmosphere*, 10, art. 642.
680 <https://doi.org/10.3390/atmos10110642>.
- 681 Diniz, A.M. 2002. Migração e evolução na fronteira agrícola. In: Anais do XIII
682 Encontro da Associação Brasileira de Estudos Populacionais, Ouro Preto, Minas
683 Gerais, 2002. Campinas, São Paulo, Brazil: Associação Brasileira de Estudos
684 Populacionais. 26 pp.
- 685 dos Santos Júnior, M.A.; Yanai, A.M.; Sousa Junior, F.O.; de Freitas, I.S.; Pinheiro,
686 H.P.; de Oliveira, A.C.R.; da Silva, F.L.; Graça, P.M.L.A.; Fearnside, P.M. 2018.
687 BR-319 como Propulsora de desmatamento: Simulando o Impacto da Rodovia
688 Manaus-Porto Velho, Instituto de Desenvolvimento Sustentável da Amazônia
689 (IDESAM), Manaus, Amazonas, Brazil. 56 pp. <https://bit.ly/3yjpdf>.
- 690 DNIT (Departamento Nacional de Infraestrutura de Transportes). 2021a. Rodovias
691 Federais. (<https://servicos.dnit.gov.br/vgeo/>). Accessed: 12 June 2021.
- 692 DNIT (Departamento Nacional de Infraestrutura de Transportes). 2021b. Rodovias
693 Estaduais. (<https://servicos.dnit.gov.br/vgeo/>). Accessed: 12 June 2021.
- 694 Dutra, D.J.; Anderson, L.O.; Fearnside, P.M.; Graça, P.M.L.d.A.; Yanai, A.M.;
695 Dalagnol, R.; Burton, C.; Jones, C.; Betts, R.; Aragão, L.E.O.e.C.d. 2023. Fire
696 dynamics in an emerging deforestation frontier in southwestern Amazonia, Brazil.
697 *Fire*, 6(2). <https://doi.org/10.3390/fire6010002>.
- 698 Fearnside, P.M. 2004. A água de São Paulo e a floresta amazônica. *Ciência Hoje*,
699 34(203), 63-65. [http://philip.inpa.gov.br/publ_livres/2004/S%20PAULO-agua-](http://philip.inpa.gov.br/publ_livres/2004/S%20PAULO-agua-C%20hoje.pdf)
700 [C%20hoje.pdf](http://philip.inpa.gov.br/publ_livres/2004/S%20PAULO-agua-C%20hoje.pdf)
- 701 Fearnside, P.M. 2015. Rios voadores e a água de São Paulo. *Amazônia Real*.
702 <https://bit.ly/3qykIsY>.
- 703 Fearnside, P.M. 2017. Deforestation of the Brazilian Amazon. In: H. Shugart (ed.)
704 *Oxford Research Encyclopedia of Environmental Science*. Oxford University
705 Press, New York, USA. <https://doi.org/10.1093/acrefore/9780199389414.013.102>.
- 706 Fearnside, P.M. 2018. BR-319 e a destruição da floresta amazônica. *Amazônia Real*, 19
707 October 2018. (<http://bit.ly/3IRS0MX>). Accessed: 14 March 2021.
- 708 Fearnside, P.M. 2021a. Deforestation in Brazilian Amazonia. In: E. Wohl (ed.) *Oxford*
709 *Bibliographies in Environmental Science*. New York, USA: Oxford University
710 Press. <https://doi.org/10.1093/obo/9780199363445-0064>.
- 711 Fearnside, P.M. 2021b. Lessons from Brazil's São Paulo droughts (commentary).
712 *Mongabay*, 30 July 2021. [https://news.mongabay.com/2021/07/lessons-from-](https://news.mongabay.com/2021/07/lessons-from-brazils-sao-paulo-droughts-commentary/)
713 [brazils-sao-paulo-droughts-commentary/](https://news.mongabay.com/2021/07/lessons-from-brazils-sao-paulo-droughts-commentary/). Accessed: 14 March 2021.

- 714 Fearnside, P.M. 2022. Amazon environmental services: Why Brazil's Highway BR-319
715 is so damaging. *Ambio*, 51, 1367–1370. [https://doi.org/10.1007/s13280-022-](https://doi.org/10.1007/s13280-022-01718-y)
716 01718-y.
- 717 Fearnside, P.M.; Ferrante, L.; Yanai, A.M.; Isaac Júnior, M.A. 2020. Trans-Purus:
718 Brazil's last intact Amazon forest at immediate risk (commentary) Mongabay, 24
719 November 2020. <https://bit.ly/3IrTJH5>. Accessed: 14 March 2021.
- 720 Fearnside, P.M.; Graça, P.M.L.A.; Keizer, E.W.H.; Maldonado, F.D.; Barbosa, R.I.;
721 Nogueira, E.M. 2009. Modelagem de desmatamento e emissões de gases de efeito
722 estufa na região sob influência da rodovia Manaus-Porto Velho (BR-319). *Revista*
723 *Brasileira de Meteorologia*, 24(2), 208-233. [https://doi.org/10.1590/S0102-](https://doi.org/10.1590/S0102-77862009000200009)
724 77862009000200009.
- 725 Fernandes, V.R.; Cunha, A.P.M.; do Amaral, P.; Luz, A.C.; Leal, K.R.D.; Costa,
726 L.C.O.; Broedel, E.; Alvalá, R.C.S.; Seluchi, M.E.; Marengo, J. 2021. Secas e os
727 impactos na região Sul do Brasil. *Revista Brasileira de Climatologia*, 28, 561-584.
728 <https://doi.org/10.5380/rbclima.v28i0.74717>.
- 729 Ferrante, L.; Fearnside, P.M. 2019. Brazil's new president and 'ruralists' threaten
730 Amazonian's environment, traditional peoples and the global climate.
731 *Environmental Conservation*, 46, 261-263.
732 <https://doi.org/10.1017/S0376892919000213>.
- 733 Franco, M.H.M. 2011. Novas configurações territoriais no Purus indígena e extrativista.
734 In: G.M. dos Santos (ed.) *Álbum Purus*. Manaus, Amazonas, Brazil: EDUA, vol. 1,
735 pp. 153-166. [https://amazonianativa.org.br/2013/10/16/novas-configuracoes-](https://amazonianativa.org.br/2013/10/16/novas-configuracoes-territoriais-no-purus-indigena-e-extrativista/)
736 [territoriais-no-purus-indigena-e-extrativista/](https://amazonianativa.org.br/2013/10/16/novas-configuracoes-territoriais-no-purus-indigena-e-extrativista/).
- 737 Gavlak, A.A. 2011. Padrões de Mudança de Cobertura da Terra e Dinâmica
738 Populacional no Distrito Florestal Sustentável da BR-163: População, Espaço e
739 Ambiente. Master's dissertation in remote sensing. São José dos Campos: INPE
740 <http://bit.ly/3kM9vX7>.
- 741 Gavlak, A.A.; Escada, M.I.S.; Monteiro, A.M.V. 2011. Dinâmica de padrões de
742 mudança de uso e cobertura da terra na região do Distrito Florestal Sustentável da
743 BR-163. In: *Simpósio Brasileiro de Sensoriamento Remoto*, 15. (SBSR), Curitiba.
744 São José dos Campos, São Paulo, Brazil: INPE, pp. 6152-6160.
745 <http://marte.dpi.inpe.br/col/dpi.inpe.br/marte/2011/06.27.12.54/doc/p1188.pdf>
- 746 Getirana, A.; Libonati, R.; Cataldi, M. 2021. Brazil is in water crisis — It needs a
747 drought plan. *Nature*, 600, 218-220. <https://doi.org/10.1038/d41586-021-03625-w>.
- 748 Hudson, W.D.; Ramm, C.W. 1987. Correct formulation of the Kappa coefficient of
749 agreement. *Photogrammetric Engineering and Remote Sensing*, 53(4), 421–422.
750 [https://www.asprs.org/wp-content/uploads/pers/1987journal/apr/1987_apr_421-](https://www.asprs.org/wp-content/uploads/pers/1987journal/apr/1987_apr_421-422.pdf)
751 422.pdf.
- 752 IBGE (Instituto Brasileiro de Geografia e Estatística). 2017. Censo Agropecuário 2017 -
753 Resultados definitivos.
754 (<https://cidades.ibge.gov.br/brasil/am/labrea/pesquisa/24/76693>). Accessed: 7
755 February 2021.
- 756 IBGE (Instituto Brasileiro de Geografia e Estatística). 2020. Solos.
757 (<http://bit.ly/3JdipGm>). Accessed: 11 January 2021.
- 758 IBGE (Instituto Brasileiro de Geografia e Estatística). 2021b. Lábrea.
759 (<https://cidades.ibge.gov.br/brasil/am/labrea/panorama>). Accessed: 13 August
760 2021.
- 761 IBGE (Instituto Brasileiro de Geografia e Estatística). 2021a. Malha Municipal.
762 (<https://bit.ly/3kQywjU>). Accessed: 14 August 2021.

- 763 ICMBIO (Instituto Chico Mendes de Conservação da Biodiversidade). 2021. Unidades
764 de Conservação. (<https://bit.ly/3kQU7bO>). Accessed: 15 January 2021.
- 765 IMAZON (Instituto do Homem e Meio Ambiente da Amazônia). 2022. Desmatamento
766 na Amazônia cresce 29% em 2021 e é o maior dos últimos 10 anos. IMAZON, 17
767 January 2022. (<http://bit.ly/3Yom4FC>). Accessed: 28 January 2022.
- 768 InfoAmazonia. 2021. Sul do Amazonas é a nova fronteira do desmatamento na
769 Amazônia, InfoAmazonia, 22 December 2021. (<http://bit.ly/3EZxqJu>). Accessed:
770 22 July 2022.
- 771 INPE (Instituto Nacional de Pesquisas Espaciais). 2020. TerraView. São José dos
772 Campos, São Paulo, Brazil: INPE. (<http://bit.ly/3JdqACK>). Accessed: 2 April
773 2022.
- 774 INPE (Instituto Nacional de Pesquisas Espaciais). 2021. Desmatamento - Amazônia
775 Legal. INPE, Coordenação Geral de Observação da Terra. Programa de
776 Monitoramento da Amazônia e Demais Biomas. (<http://bit.ly/3ZuuFIw>). Accessed:
777 9 June 2021.
- 778 IPAM (Instituto de Pesquisa Ambiental da Amazônia). 2021. Invasão de terras públicas
779 foi a principal causa do desmatamento na Amazônia. IPAM, 22 November 2021.
780 (<http://bit.ly/3ZjFoFo>). Accessed: 3 January 2022.
- 781 Körting, T.S.; Fonseca, L.M.G.; Escada, M.I.S.; Silva, F.C.; Silva, M. 2008. GeoDMA
782 – A novel system for spatial data mining. In: ICDMW '08: Proceedings of the 2008
783 IEEE International Conference on Data Mining Workshops, December 2008. Pisa,
784 Italy. pp. 975–978. <https://doi.org/10.1109/ICDMW.2008.22>.
- 785 Laurance, W.F.; Cochrane, M.A.; Bergen, S.; Fearnside, P.M.; Delamônica, P.; Barber,
786 C.; D'Angelo, S.; Fernandes, T. 2001. The future of the Brazilian Amazon.
787 *Science*, 291, 438-439. <https://doi.org/10.1126/science.291.5503.438>.
- 788 Marinaro, S.; Sacchi, L.; Gasparri, N.I. 2022. From whom and for what: Deforestation
789 in Dry Chaco from local-urban inhabitants' perceptions. *Perspectives in Ecology
790 and Conservation*, 20, 141-150. <https://doi.org/10.1016/j.pecon.2021.12.003>.
- 791 Maurano, L.E.P.; Escada, M.I.S.; Renno, C.D. 2019. Padrões espaciais de
792 desmatamento e a estimativa da exatidão dos mapas do PRODES para Amazônia
793 Legal Brasileira. *Ciência Florestal [Santa Maria]*, 29(4), 1763-1775.
794 <https://doi.org/10.5902/1980509834380>.
- 795 McGarigal, K.; Marks, B.J. 1995. FRAGSTATS: Spatial pattern analysis program for
796 quantifying landscape structure. Washington: United States Department of
797 Agriculture, 132 pp. (Pacific Northwest Research Station General Technical
798 Report PNW-GTR-351). https://www.fs.usda.gov/pnw/pubs/pnw_gtr351.pdf.
- 799 Mertens B.; Lambin, E.F. 1997. Spatial modeling of deforestation in southern
800 Cameroon. Spatial disaggregation of diverse deforestation processes. *Applied
801 Geography*, 17(2), 143-162. [https://doi.org/10.1016/S0143-6228\(97\)00032-5](https://doi.org/10.1016/S0143-6228(97)00032-5).
- 802 Moutinho, P. 2022. Comissão de Meio Ambiente debate denúncia fraude no Cadastro
803 Ambiental Rural em 25 de maio de 2022. TV Senado. (<http://bit.ly/3KZh07D>).
804 Accessed: 2 May 2022.
- 805 NICFI (Norway's International Climate and Forest Initiative). 2021. NICFI Data
806 Program. <https://www.nicfi.no/>.
- 807 Nicolini, M.; Saulo, C.; Torres, J.C.; Salio, P. 2002. Enhanced precipitation over
808 southeastern South America related to low-level jet events during austral warm
809 season. *Meteorologica*, 27, 59–70.
810 [https://ri.conicet.gov.ar/bitstream/handle/11336/148157/CONICET_Digital_Nro.d
811 d6add01-8def-48e5-80f7-250a4f9f98af_A.pdf?sequence=2&isAllowed=y](https://ri.conicet.gov.ar/bitstream/handle/11336/148157/CONICET_Digital_Nro.d6add01-8def-48e5-80f7-250a4f9f98af_A.pdf?sequence=2&isAllowed=y).

- 812 Nobre, C., Marengo, J., Seluchi, M., Cuartas, L.; Alves, L. 2016. Some characteristics
813 and impacts of the drought and water crisis in southeastern Brazil during 2014 and
814 2015. *Journal of Water Resource and Protection*, 8, 252-262.
815 <https://doi.org/10.4236/jwarp.2016.82022>.
- 816 Nogueira, E.M.; Yanai, A.M.; Fonseca, F.O.R.; Fearnside, P.M. 2015. Carbon stock
817 loss from deforestation through 2013 in Brazilian Amazonia. *Global Change*
818 *Biology* 21, 1271–1292. <https://doi.org/10.1111/gcb.12798>.
- 819 Oliveira Filho, F.J.B.; Metzger, J.P. 2006. Threshold in landscape structure for three
820 common deforestation patterns in the Brazilian Amazon. *Landscape Ecology*, 21,
821 1061-1073. <https://doi.org/10.1007/s10980-006-6913-0>.
- 822 Pack, S. M.; Ferreira, N.F.; Krithivasan, R.; Murrow, J.; Bernard, E.; Mascia, M. B.
823 2016. Protected area downgrading, downsizing, and degazettement (PADD) in the
824 Amazon. *Biological Conservation*, 197, 32-39.
825 <https://doi.org/10.1016/j.biocon.2016.02.004>.
- 826 Pereira, L.M.; Escada, M.I.; Rennó, C.D. 2007. Análise da evolução do desmatamento
827 em áreas de pequenas, médias e grandes propriedades na região centro-norte de
828 Rondônia, entre 1985 e 2000. *Anais do XIII Simpósio Brasileiro de Sensoriamento*
829 *Remoto*, Florianópolis, Brasil, 21-26 abril 2007, INPE, pp. 6905-6912.
830 [http://mar.te.sid.inpe.br/col/dpi.inpe.br/sbsr@80/2006/11.10.12.31/doc/6905-](http://mar.te.sid.inpe.br/col/dpi.inpe.br/sbsr@80/2006/11.10.12.31/doc/6905-6912.pdf)
831 [6912.pdf](http://mar.te.sid.inpe.br/col/dpi.inpe.br/sbsr@80/2006/11.10.12.31/doc/6905-6912.pdf).
- 832 Pinheiro, T.F.; Escada, M.I.S.; Valeriano, D.M.; Hostert, P.; Gollnow, F.; Müller, H.
833 2016. Forest degradation associated with logging frontier expansion in the
834 Amazon: The BR-163 region in southwestern Pará, Brazil. *Earth Interactions*, 20,
835 art. 17. <https://doi.org/10.1175/EI-D-15-0016.1>.
- 836 Quinlan, J.R. 2017. C5.0: An Informal Tutorial. (<https://bit.ly/3ygRYtk>). Accessed: 9
837 May 2021.
- 838 RADAMBRASIL 1978. Folha no. SB 20 Purus: geologia, pedologia, vegetação, e uso
839 potencial da terra. Rio de Janeiro, RJ, Brazil: Departamento Nacional de Produção
840 Mineral. 566 pp. [https://biblioteca.ibge.gov.br/index.php/biblioteca-](https://biblioteca.ibge.gov.br/index.php/biblioteca-catalogo?view=detalhes&id=23596)
841 [catalogo?view=detalhes&id=23596](https://biblioteca.ibge.gov.br/index.php/biblioteca-catalogo?view=detalhes&id=23596).
- 842 Saito, E.A. 2010. Caracterização de Trajetórias de Padrões de Ocupação Humana na
843 Amazônia Legal por Meio de Mineração de Dados.
844 (<http://urlib.net/8JMKD3MGP7W/38MM2TL>). Master's dissertation in remote
845 sensing, Instituto Nacional de Pesquisas Espaciais, São José dos Campos, SP. 158
846 pp. (<https://bit.ly/3YouQn5>). Accessed: 12 December 2020.
- 847 Saito, E.A.; Fonseca, L.M.G.; Escada, M.I.S.; Körting, T.S. 2011. Análise de padrões de
848 desmatamento e trajetória de padrões de ocupação humana na Amazônia usando
849 técnicas de mineração de dados. In: *Anais XV Simpósio Brasileiro de*
850 *Sensoriamento Remoto - SBSR*, Curitiba, PR, Brasil, 30 de abril a 05 de maio de
851 2011, São José dos Campos, São Paulo, Brazil: INPE. pp. 2833-2840.
852 <https://bit.ly/3ST0eZP>.
- 853 Saito, E.A.; Fonseca, L.M.G.; Escada, M.I.S.; Körting, T.S. 2012. Efeitos da mudança
854 de escala em padrões de desmatamento na Amazônia. *Revista Brasileira de*
855 *Cartografia*, 63(3), 401–414. <https://doi.org/10.14393/rbcv63n3-43749>.
- 856 Sampaio, L.S.; Costa, R.G.C. 2009. Estradas e suas relações socioambientais. In: Porro,
857 R. (ed.). *Alternativa Agroflorestal na Amazônia em Transformação*. Brasília, DF,
858 Brazil: Embrapa Informação Tecnológica, 825 pp.
859 [https://www.worldagroforestry.org/publication/alternativa-agroflorestal-na-](https://www.worldagroforestry.org/publication/alternativa-agroflorestal-na-amazonia-em-transformacao)
860 [amazonia-em-transformacao](https://www.worldagroforestry.org/publication/alternativa-agroflorestal-na-amazonia-em-transformacao).

- 861 Santos, J.L.; Yanai, A.M.; Graça, P.M.L.A.; Correia, F.W.S.; Fearnside, P.M. 2023.
862 Amazon deforestation: Simulated impact of Brazil's proposed BR-319 highway
863 project. *Environmental Monitoring and Assessment* 195(10): art. 1217.
864 <https://doi.org/10.1007/s10661-023-11820-7>.
- 865 Schubart, H.O.R. 1983. Ecologia e utilização das florestas. In: Salati, E. *Amazônia:*
866 *Desenvolvimento, Integração e Ecologia*. São Paulo, Brazil: Brasiliense, pp. 101-
867 143.
- 868 Silva, C.A.; Santilli, G.; Sano, E.E.; Laneve, G. 2021. Fire occurrences and greenhouse
869 gas emissions from deforestation in the Brazilian Amazon. *Remote Sensing*,
870 13(376). <https://doi.org/10.3390/rs13030376>.
- 871 Silva, F.C.; Körting, T.S.; Fonseca, L.M.G.; Escada, M.I.S. 2007. Deforestation pattern
872 characterization in the Brazilian Amazonia. In: *Anais do XIII Simpósio Brasileiro*
873 *de Sensoriamento Remoto*, 13. (SBSR), Florianópolis. São José dos Campos, São
874 Paulo, Brazil: INPE. pp. 6207-6214. <https://bit.ly/3kS9lgY>.
- 875 Silva, M.P.S.; Câmara, G.; Escada, M.I.S.; Souza, R.C.M. 2008. Remote-sensing image
876 mining: detecting agents of land-use change in tropical forest areas. *International*
877 *Journal of Remote Sensing*, 29(16), 4803-4822.
878 <https://doi.org/10.1080/01431160801950634>.
- 879 Soares-Filho, B.; Moutinho, P.; Nepstad, D.; Anderson, A.; Rodrigues, H.; Garcia, R.
880 2010. Role of Brazilian Amazon protected areas in climate mitigation. In: 24 (Ed),
881 *Proceedings of the National Academy of Sciences USA*, 107, 10821–10826.
882 <http://www.pnas.org/cgi/doi/10.1073/pnas.0913048107>.
- 883 Souza, A.A.A.; Pontes, A.N.; Adami, M.; Narvaes, I.S. 2017. A contribuição das
884 estradas e o padrão de desflorestamento e degradação da cobertura florestal no
885 sudoeste paraense. *Revista Brasileira de Cartografia*, 69(9), 1711-1724.
886 <https://doi.org/10.14393/rbcv69n9-44089>.
- 887 SUDAM (Superintendência do desenvolvimento da Amazônia). 2021. Zona de
888 desenvolvimento sustentável dos Estados do Amazonas, Acre e Rondônia 2021-
889 2027: documento referencial. Belém, Pará: SUDAM. 174 pp. [https://irp.cdn-](https://irp.cdn-website.com/06e645c5/files/uploaded/amacro.pdf)
890 [website.com/06e645c5/files/uploaded/amacro.pdf](https://irp.cdn-website.com/06e645c5/files/uploaded/amacro.pdf).
- 891 Tavares, L.; Cordeiro, L. 2017. Perfil Socioeconômico e Ambiental do Sul do Estado do
892 Amazonas: Subsídios para Análise da Paisagem. Brasília, DF, Brazil: WWF–
893 Brasil. 56 pp. <http://bit.ly/3kQVuHu>.
- 894 Tomasi, A.S. 2018. Grilagem de Terras e Violência Agrária: Criação e Extinção de um
895 Projeto de Reforma Agrária no sul do Amazonas. Campinas, São Paulo, Brazil:
896 Grupo Governança de Terras, Universidade Estadual de Campinas (UNICAMP),
897 45 pp. <https://bit.ly/3mjBYno>.
- 898 Vale, M.M.; Berenguer, E.; de Menezes, M.A.; de Castro, E.B.V; de Siqueira, L.P.;
899 Portela, R.C.Q. 2021. The COVID-19 pandemic as an opportunity to weaken
900 environmental protection in Brazil. *Biological Conservation*, 255, art. 108994.
901 <https://doi.org/10.1016/j.biocon.2021.108994>.
- 902 van der Ent, R. J.; Savenije, H. H. G.; Schaeffli, B.; Steele-Dunne, S. C. 2010. Origin
903 and fate of atmospheric moisture over continents. *Water Resources Research*, 46,
904 W09525, <https://doi.org/10.1029/2010WR009127>.
- 905 Vitel, C.S.M.N; Fearnside, P.M.; Graça, P.M.L.A. 2009. Análise da inibição do
906 desmatamento na parte Sudoeste do Arco de desmatamento. In: J.C.N. Epiphany
907 & L.S. Galvão (Eds.) *Anais XIV Simpósio Brasileiro de Sensoriamento Remoto*,
908 Natal, Brasil 2009. São José dos Campos, São Paulo, Brazil: Instituto Nacional de
909 Pesquisas Espaciais (INPE), pp. 6377-6384. <https://bit.ly/3EUzORx>.

- 910 Walker, W.S.; Gorelik, S.R.; Bazzini, A.; Aragon-Osejo, J.L.; Josse, C.; Meyer, C.; et
911 al. 2020. The role of forest conversion, degradation, and disturbance in the carbon
912 dynamics of Amazon indigenous territories and protected areas. *Proceedings of the*
913 *National Academy of Sciences USA*, 117(6), 3015-3025.
914 <https://doi.org/10.1073/pnas.1913321117>.
- 915 Wenzel, F. 2023. Os cupins da floresta: Com Bolsonaro, Ministério do Meio Ambiente
916 abriu mão de área na Amazônia onde madeireiros derrubam 45 mil caminhões de
917 árvores. *The Intercept Brasil*, 17 January 2023. (<http://bit.ly/3JGysNy>) Accessed:
918 18 January 2023.
- 919 Yanai, A.M.; Nogueira, E. M.; Graça, P.M.L.A.; Fearnside, P.M. 2017. Deforestation
920 and carbon stocks loss in Brazil's Amazonian settlements. *Environmental*
921 *Management*, 59, 393-409. <https://doi.org/10.1007/s00267-016-0783-2>.
- 922 Yanai, A.M.; Graça, P.M.L.A.; Escada, M.I.S.; Ziccardi, L.G.; Fearnside, P.M. 2020.
923 Deforestation dynamics in Brazil's Amazonian settlements: Effects on land-tenure
924 concentration. *Journal of Environmental Management*, 268, art. 110555.
925 <https://doi.org/10.1016/j.jenvman.2020.110555>.
- 926 Yanai, A.M.; Graça, P.M.L.A.; Ziccardi, L.G.; Escada, M.I.S.; Fearnside, P.M. 2022.
927 Brazil's Amazonian deforestation: The role of landholdings in undesignated public
928 lands. *Regional Environmental Change*, 22, art. 30.
929 <https://doi.org/10.1007/s10113-022-01897-0>.
- 930 Zemp, D.C., Schleussner, C.-F.; Barbosa, H.M.J.; van der Ent, R.J.; Donges, J.F.;
931 Heinke, J.; Sampaio, G.; Rammig, A. 2014. On the importance of cascading
932 moisture recycling in South America. *Atmospheric Chemistry and Physics*, 14,
933 13.337–13.359. <https://doi.org/10.5194/acp-14-13337-2014>.

Supplementary Material

Amazon deforestation: A dangerous future indicated by patterns and trajectories in a hotspot of forest destruction in Brazil

Beatriz Figueiredo Cabral¹, Aurora Miho Yanai¹, Paulo Maurício Lima de Alencastro Graça¹, Maria Isabel Sobral Escada², Cláudia Maria de Almeida², Philip Martin Fearnside¹

¹ Instituto Nacional de Pesquisas da Amazônia (INPA), Av. André Araújo n° 2936, CEP 69067-375, Manaus, Amazonas, Brazil, beatriz.figueiredocabral@gmail.com; yanai@inpa.gov.br pmlag@inpa.gov.br; pmfearn@inpa.gov.br

² Instituto Nacional de Pesquisas Espaciais (INPE), Av. dos Astronautas, 1758, CEP 12227-010, São José dos Campos, São Paulo, Brazil, isabel.escada@inpe.br; claudia.almeida@inpe.br

Table S.1. Explanation of Landscape Metrics-Based features. (Adapted from McGarigal and Marks, 1995)

Name	Description	Formula
AWMSI	The Area-Weighted Mean Shape Index (AWMSI) is the ratio of the perimeter to the square root of the area, weighted by the patch area, so that larger patches carry more weight than smaller ones, with $AWMSI \geq 1$. $AWMSI = 1$ if all patches are circular or rectangular. This index increases as the complexity of polygons rises, meaning that it increases as shapes become more irregular.	$AWMSI = \sum_{j=1}^n \left[\frac{p_j}{2\sqrt{\pi \times a_j}} \times \frac{a_j}{\sum_{j=1}^n a_j} \right]$
AWMPDF	Area-Weighted Mean Patch Fractal (AWMPDF) stands for the fractal dimension of the patch, weighted by the patch area in the landscape. $1 \leq AWMPDF \leq 2$, measures the irregularity or complexity of the patch shape. AWMPDF approaches 1 for patches with simpler shapes, such as circles or rectangles, and approaches 2 when patches have a more complex shape.	$AWMPDF = \sum_{j=1}^n \left[\frac{2 \times \ln p_j}{\ln a_j} \times \frac{a_j}{\sum_{j=1}^n a_j} \right]$
BIA	BIA stands for Biggest Intersection Area of the class in the landscape.	-
CA	Class Area (CA) is a measure of landscape composition represented by the sum of the areas of all fragments of a specific class in hectares (ha). $CA > 0$. The CA value approaches 0 when there are few patches of the class in the landscape.	$CA = \sum_{j=1}^n a_j$

LSI	The Landscape Shape Index (LSI) measures the shape complexity of patches, with $LSI \geq 1$, with no set limit. $LSI = 1$ when the landscape consists of a single patch with a circular or rectangular shape.	$LSI = \frac{\sum_{j=1}^n e_j}{2\sqrt{\pi \times A}}$
PD	PD stands for Patch Density, which equals the number of patches in the landscape divided by total landscape area (m^2), multiplied by 10,000 and 100 (to convert to 100 hectares). Note, PD does not include background patches or patches in the landscape border, if present. However, total landscape area (A) includes any internal background present.	$PD = \frac{n}{A} \times 10000 \times 100$
PR	PR stands for Patch Richness, which equals the number of different patch types (classes) present within the landscape boundary.	$PR = m$
TABO	TABO stands for the Total Area of the Biggest Object that intersects the landscape.	-

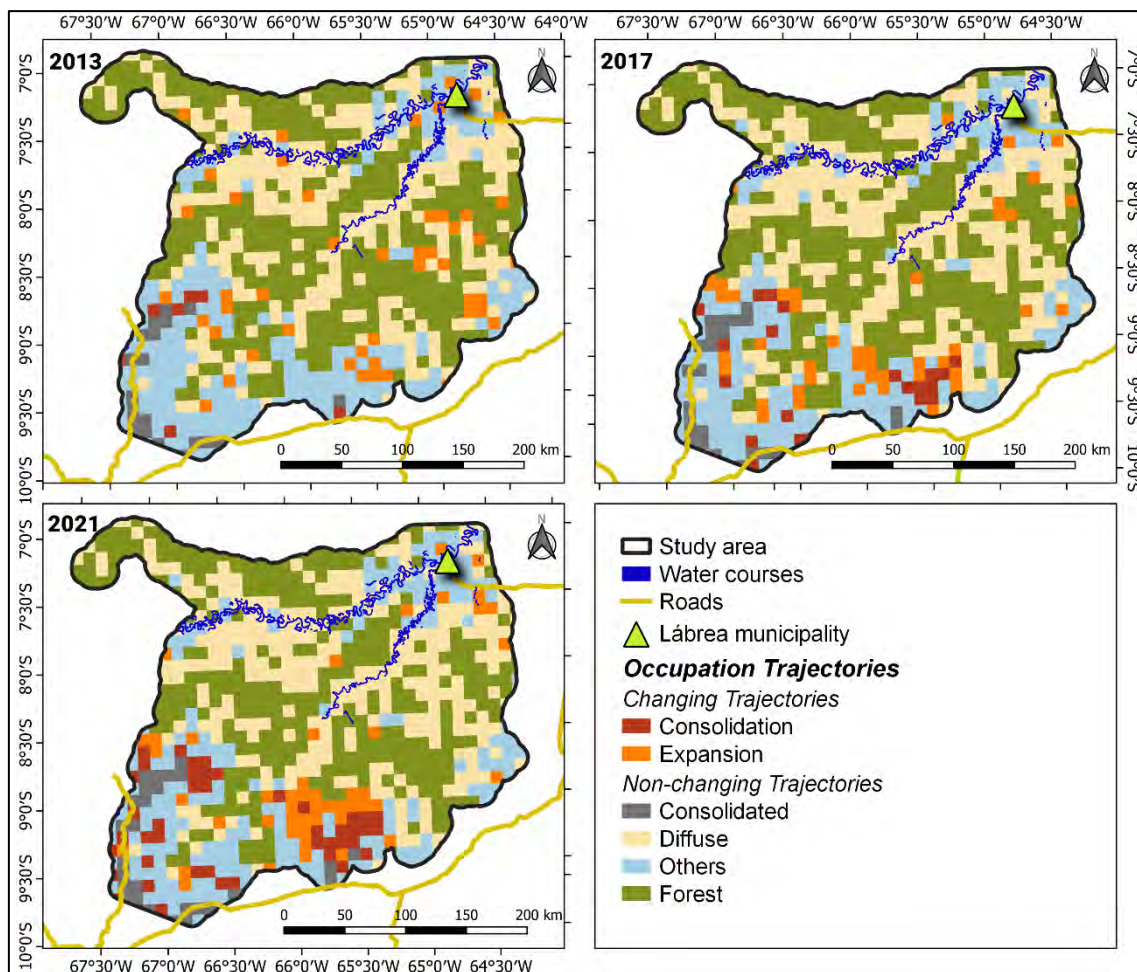


Figure S.1. Maps of the evolution of occupation trajectories in the study area for the periods 2008 to 2013, 2013 to 2017, and 2017 to 2021.

Table S.2. Percentage of deforested cells and total deforested area (in hectares) for each deforestation trajectory from 2008 to 2021 in the study area.

Deforestation trajectory	Percentage of cells (%)	Deforested area (ha)	Percentage of total deforested area (%)
Consolidation	8.2%	302,297.19	30.8%
Expansion	21.2%	190,824.69	19.6%
Consolidated	3.6%	170,081.82	17.4%
Diffuse	41.2%	13,609.84	1.4%
Others			
Fishbone	3.8%	59,238.17	6.0%
Geometric	5.6%	113,078.94	11.6%
Linear	13.6%	31,718.55	3.3%
Multidirectional	2.8%	96,376.84	9.9%
Subtotal	25.8%	300,412.50	30.8%
TOTAL	100%	977,226.04	100%

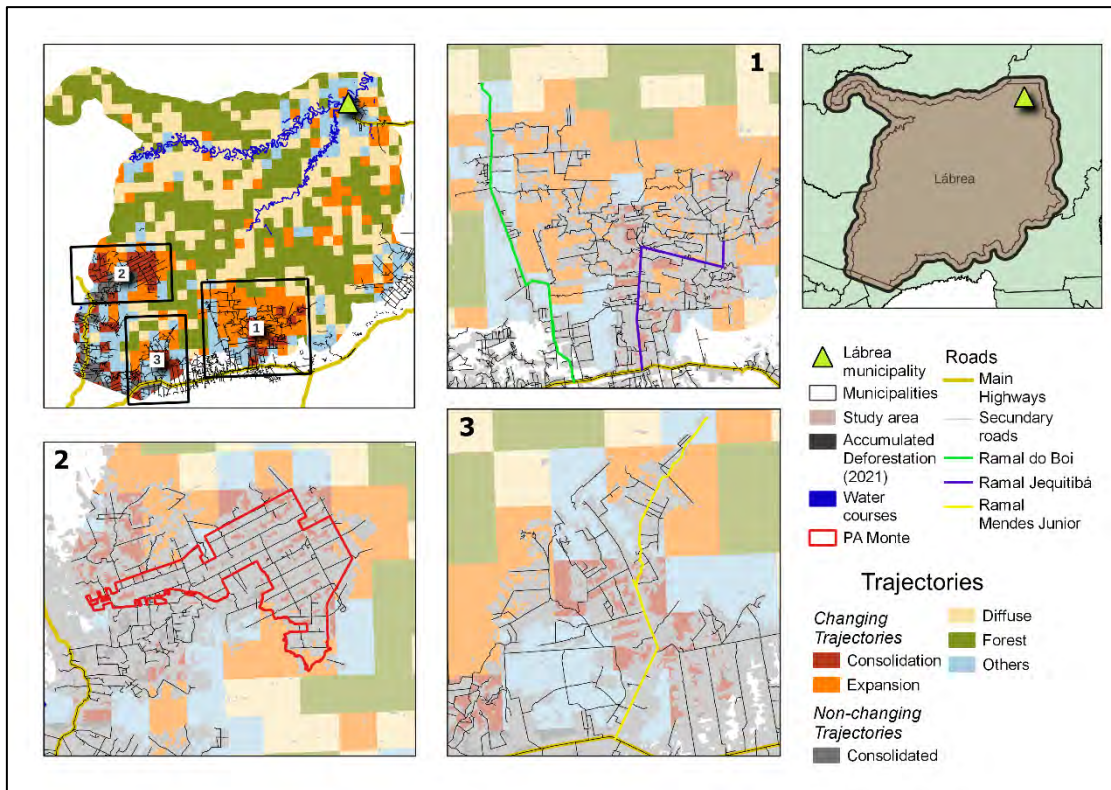


Figure S.2. Areas with advance of the "consolidation" and "expansion" change trajectories for the study area (i.e., the municipality of Lábrea + 10-km buffer) from 2008 to 2021.

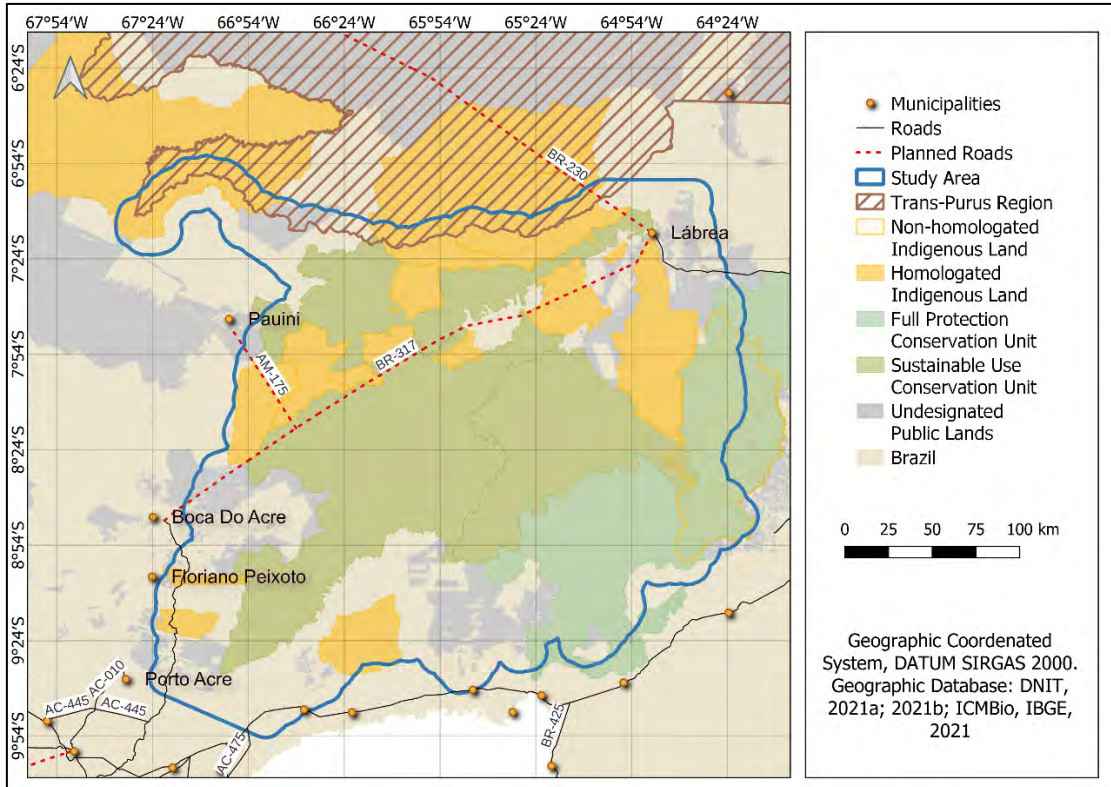


Figure S.3. Map of planned state and federal highways for the municipality of Lábrea, Amazonas.

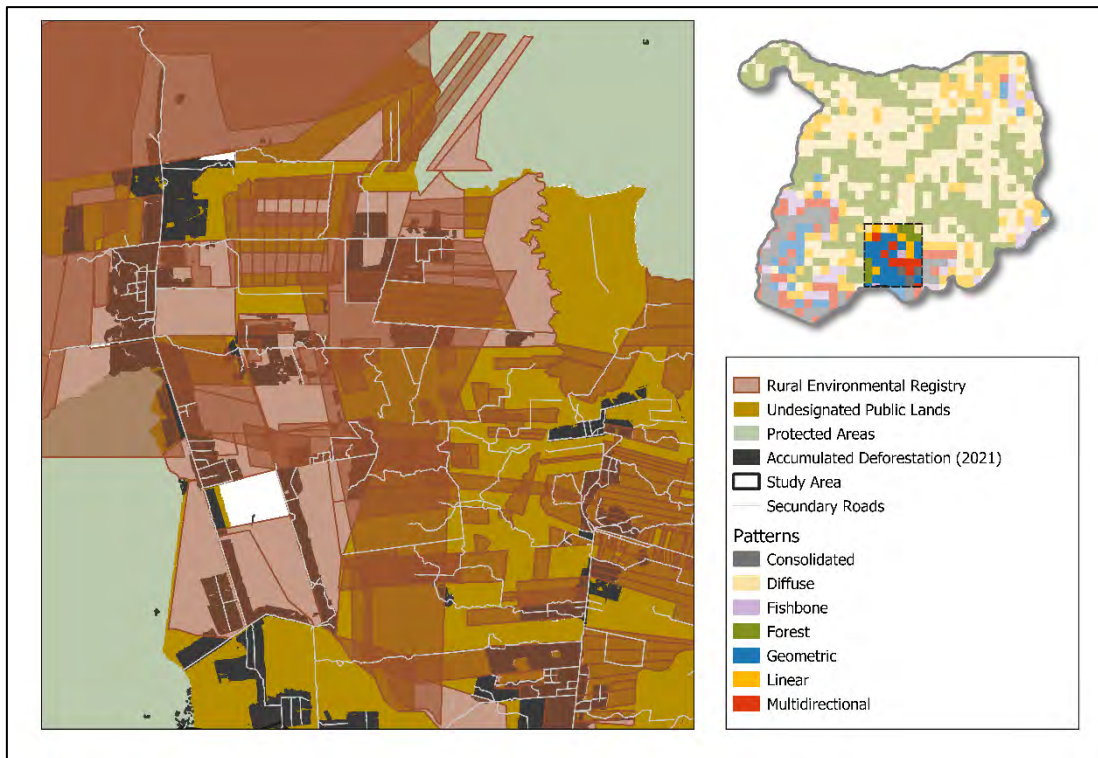


Figure S.4. Rural Environmental Registry (CAR) claims overlapping undesignated public lands in 2021, showing concentrations of the geometric deforestation pattern.

Reference

McGarigal, K.; Marks, B.J. 1995. FRAGSTATS: Spatial pattern analysis program for quantifying landscape structure. Washington: United States Department of Agriculture, 132 pp. (Pacific Northwest Research Station General Technical Report PNW-GTR-351). https://www.fs.usda.gov/pnw/pubs/pnw_gtr351.pdf